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1 Geogenic arsenic and uranium in Germany: large-scale distribution control in sediments

2 and groundwater

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Abstract

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Arsenic (As) and uranium (U) are naturally occurring trace elements with potentially adverse effects on human health. This work revisits nine case studies on As/U accumulation and remobilization mechanisms in aquifers with different geological and stratigraphical backgrounds to develop a systematic overview of Germany's geogenic inventory of these trace elements. It uses geochemical proxies for a total of 270 solid samples to explain their spatio-temporal distribution: while Pleistocene geological development can explain their extensive absence in sediments and related groundwater in northern Germany, their abundance and distribution in the central and southern parts are widely controlled by sediment provenance geochemistry. Only highly felsic origin (Moldanubian Variscides) enables creation of elevated U in the systems while lower degrees of provenance felsicity (Rhenohercynian Variscides) appear to be sufficient for As presence. Postdepositional (hydro)geological and anthropogenically triggered intra-basinal processes of trace element accumulation, redistribution and eventually remobilization to groundwater contibute to the present-day situation. Therefore, the ultimate control of these incompatible trace elements is magmatic, even in old sedimentary systems, and still clearly traceable in nowadays large-scale geogenic As and U distribution in Germany and probably elsewhere.

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- 32 Keywords
- 33 magmatic differentiation, geochemical proxies, sediment provenance, Variscides, accumula-
- 34 tion processes

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1 Introduction

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1.1 Arsenic and uranium – a short geochemical overview

- 40 Showing characteristics of both chalcophile and siderophile behaviour, arsenic (As) tends to be prefer-41 ably hosted by sulphide minerals like pyrite or (hydr)oxidic Fe phases like goethite, both of which can 42 contain As up to several wt.% (Smedley and Kinniburgh 2002). In spite of its low average abundance in the upper earth's crust (1.5-2 µg g⁻¹; Matschullat 1999, Taylor 1964: 1.8 µg g⁻¹ in the continental 43 crust). As can accumulate in rocks to concentrations several orders of magnitude higher than these 44 45 values. The metalloid's fate in the environment is controlled by the prevailing physico-chemical con-46 ditions and the presence of other ions. Redox milieu, pH and ionic competition are crucial parameters 47 governing As behaviour (adsorption, desorption, transport, redox transformation). Reducing condi-48 tions can lead to As mobilization from oxides while oxidizing conditions may mobilize As bound to 49 sulphides. High groundwater pH constrains As adsorption to mineral surfaces and may therefore be 50 responsible for elevated concentrations in solution. Ions competing with As species for surface binding 51 sites, especially phosphate, can lead to the same result (e.g. Stollenwerk 2002, Smedley and Kinni-52 burgh 2002, Pedersen et al. 2006).
- The mobility of U in the environment is, akin to that of As, governed by the Eh-pH milieu and the
- 54 presence of adsorbers like Fe (hydr)oxides, clay minerals or organic matter (e.g. Doi et al. 1975,
- 55 Giblin et al. 1981, Merkel and Sperling 1998, Missana et al. 2003, Bots and Behrends 2008). Never-
- 56 theless, contrary to As, U is significantly more mobile in its oxidized form U(VI) which is reasoned by
- 57 its affinity to form stable uranyl hydroxo or calcium uranyl carbonato complexes (e.g. Katsoyiannis et
- al. 2007, Stewart et al. 2010). In spite of their differing redox and transport properties, As and U fre-
- 59 quently occur together in (ground)water of affected areas (e.g. Brown et al. 2007, Kipp et al. 2009,
- Nicolli et al. 2010, Banning et al. 2012, Banning and Rüde 2015).
- Taylor (1964) cites the average uranium (U) abundance in the continental crust as 2.7 µg g⁻¹. Thereby,
- 62 generally higher concentrations are detected in felsic rocks (granite average: 4.8 μg g⁻¹ U) compared to
- mafic lithologies (basalt average: 0.6 µg g⁻¹ U). Uranium has a strongly incompatible behavior in sili-
- cate magmatic differentiation because of its large ion radius and high valence. It is preferentially
- 65 fractionated into high-temperature metaluminous melts during partial melting and crystal fractionation
- 66 (Cuney 2010), and therefore often accumulates in granitic and pegmatitic lithologies. Also As is con-
- sidered one of the incompatible elements which do not fit easily into the lattices of rock-forming
- 68 minerals precipitating from melt (Webster and Nordstrom 2003).

1.2 Arsenic and uranium as contaminants – health aspects and affected areas

- 70 The recognition of As toxicity and its initially underestimated impact on human health on a global
- scale substantially increased the intensity of As research during the past decades, focusing on large
- 72 problem areas, especially in Southeast Asia (e.g. Berg et al. 2001, Ravenscroft et al. 2005, Zahid et al.

73 2009). Consequently, drinking water threshold values were broadly lowered in the 1990ies, mostly 74 down to 10 µg L⁻¹. This confronted water suppliers with the problem of an increased need to process 75 raw water in order to match drinking water requirements. In recent years, it was also found that in the 76 large majority of cases, naturally occurring As is responsible for elevated groundwater concentrations. 77 Thereby, mobilization from As-enriched minerals is the dominating process (e.g. Lowers et al. 2007). 78 Besides the large tropical, mainly deltaic regions with As-exposed populations (e.g. Bangladesh, West 79 Bengal, Vietnam, Taiwan), (semi)arid As problem areas have been identified worldwide. Prominent 80 examples can be found in Chile (e.g. Oyarzun et al. 2004) and the western U.S.A. (e.g. Welch and 81 Lico 1998). Identified As problem areas in Mexico include the Zimapán Valley (Armienta et al. 2001), 82 the Rioverde Basin (Planer-Friedrich et al. 2001) and the Villa de Reyes Graben around San Luis Po-83 tosí (Banning et al. 2012). Even endemic As poisoning was described in the Lagunera region (Del 84 Razo et al. 1990). An overview of elevated As occurences in Mexico is given by Armienta and Sego-85 via (2008). 86 While the toxicity of As is well documented and drinking water limitations are established and recon-87 sidered for several decades (actually 10 µg L⁻¹, WHO 2006), U was neglected in this respect for a long 88 time. It was shown that the risk of U exposure primarily derives from its toxicity as a nephrotoxic 89 heavy metal (i.e., leading to kidney deseases), rather than from its radioactive character (Zamora et al. 90 1998, Orloff et al. 2004, Bjørklund et al. 2017). For adults in Germany, a total radiation exposure of 91 2.1 mSv a⁻¹ was determined, whereby exposure via drinking water only contributes 0.009 mSv a⁻¹ (BfS 92 2009), i.e. ~0.4 % of the annual radiation dosage. There is no general agreement on fixed limitations 93 for U concentrations in drinking water up to date, although drinking water is considered the most im-94 portant source of U uptake. WHO announced a "provisional guideline value" of 30 µg L⁻¹, German 95 legislation decided on a fixed limitation of 10 µg L⁻¹, valid since 2011. Compared to As, the number of 96 identified geogenic groundwater U problem areas worldwide is low but growing, Nevertheless, it has 97 been recognized that high-U aquifers represent a phenomenon of global extent. Felsic magmatic aqui-98 fers in Scandinavia are partly affected by elevated concentrations (Frengstad et al. 2000 and citations 99 therein). Sherman et al. (2007) found 29 % of the studied water samples from a sandstone aquifer in 100 Michigan in excess of the WHO guideline. These authors also give a short overview of case studies 101 from other affected states and countries. In the course of a national monitoring of domestic well water 102 in the U.S.A., Focazio et al. (2006) detected 4 % of all samples exceeding 30 $\mu g \ L^{-1} \ U$. In the north-103 west Indian state of Punjab, high U concentrations in groundwater have relatively recently been ob-104 served over a wide geographical area. Several studies (Patnaik et al. 2015; Kumar et al. 2016; Bajwa et 105 al. 2017; Lapworth et al. 2017) point to potential risks for the exposed population. Unverified hypoth-106 eses on the origin of groundwater U include fly ash from coal combustion or input from agricultural 107 fertilisation. However, most authors consider geogenic processes more likely, primarily the weathering 108 of the Siwaliks, i.e. Himalayan foothills (e.g., Patnaik et al. 2015). Also high As groundwaters are 109 known from the region (e.g., Kumar and Singh 2020).

1.3 Known cases of elevated As and U contents in German sediments and groundwater

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111 While the mechanism of large-scale As accumulation in delta sediments of Southeast Asia (e.g. 112 Acharyya et al. 2000, Stanger 2005) or in (semi)arid Latin America (e.g. Bundschuh et al. 2004, 113 Nicolli and Bundschuh 2010) are subject to extensive research and discussions, relatively little was 114 known about the origin and development of geogenic As accumulations in Germany. These have been 115 studied on a rather local to regional scale, although elevated As concentrations in groundwater have 116 been detected in many parts of the country, an overview is given by Heinrichs and Udluft (1996). Fig-117 ure 1 illustrates a survey of conducted studies on elevated As occurrences in sedimentary rocks and 118 partly groundwater, whereby no claim to completeness is made. It does neither include high As ob-119 served in thermal waters known from e.g., Baden-Baden (Rüde 1996), Aachen (Herch 1997) and 120 Wiesbaden (Rosenberg et al. 1999), nor anthropogenically induced As problems (e.g., from former 121 mining, past agricultural techniques or ammunition). The As-related studies shown in Fig. 1 will be 122 shortly summarized in the following. 123 Banning et al. (2009) detected partly high As contents up to 140 µg g⁻¹ in secondary Fe concretions in 124 oxidized Upper Cretaceous (Santonian) sediments from the Münsterland Cretaceous Basin (area 2 in 125 Fig. 1) and identified them as the source for significant As contents in soils of the region. A postdepo-126 sitional paleo redox event during the late Tertiary caused extremely heterogeneous As and other trace 127 element distribution in shallow marine sediments, leading to partly massively enriched secondary con-128 cretions. Mainly pyrite-controlled, rather homogeneous As distribution in the original deeper sediment 129 facies changes under formation of highly reactive redox transition zones and distinct paleo redox 130 boundaries into extremely heterogeneous Fe hydroxide-controlled distribution in near-surface sedi-131 ments (Banning et al. 2013a). The same mechanism applied to Tertiary marine sediments in the Lower 132 Rhine Embayment (area 5; Banning and Rüde 2010): Santonian and Chattian shallow marine sedi-133 ments exhibit very similar responses to the late Tertiary oxidative redox event, including rock fabric 134 and mineralogical changes, trace element (esp. As) redistribution and remobilization potentials. The 135 developed genetic and geochemical patterns are obviously of general validity for the studied geo-136 environment: large-scale redox events alter comparable lithologies and redistribute trace elements 137 hosted therein in the same way, independent of sediment age. 138 Concentrations of As in groundwater from southern Lower Saxony (area 3) above the drinking water 139 standard were attributed to output from Lower Triassic ("Buntsandstein") clastic sediments (up to 693 140 μg g⁻¹ As) by Goldberg et al. (1995). Mertens (2000) found an average of 108 μg g⁻¹ As in Upper Cre-141 taceous (Cenomanian) glauconitic sands in the Ruhr Area (area 4) whereby output from the rocks to 142 groundwater was not observed. In a Pliocene aquifer from the Lower Rhine Embayment (area 5), 143 Cremer et al. (2003) detected up to 130 µg L⁻¹ As in groundwater, attributable to mobilization via py-144 rite oxidation, probably triggered by anthropogenic NO₃ input.

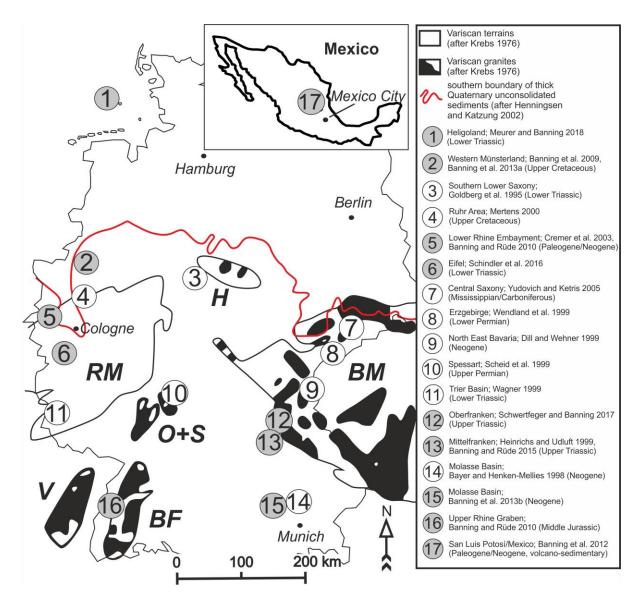


Fig. 1: Distribution of case studies on geogenic As and U in sedimentary rocks and groundwater in Germany (and one study in Mexico; numbering ascending from north to south; studies considered in this work marked as filled grey circles), Variscan terrains and granite intrusions therein (after Krebs 1976) and southern boundary of thick Quaternary cover sediments (after Henningsen and Katzung 2002). Variscan terrains from north to south: H: Harz Mountains, RM: Rhenish Massif, BM: Bohemian Massif, O+S: Odenwald and Spessart, BF: Black Forest, V: Vosges.

Arsenic contents are partly elevated in coal, an overview is given by Yudovich and Ketris (2005). Therein, the authors mention contents of up to $400 \ \mu g \ g^{-1}$ As in Mississippian (Lower Carboniferous) bituminous coals from Saxony (area 7). Interestingly, they found that Eastern German black coal is significantly more enriched than the Western German Upper Carboniferous coal with average As contents of $6.8 \ \mu g \ g^{-1}$. For comparison: the Clarke value for bituminous coal is $9.0 \pm 0.8 \ \mu g \ g^{-1}$. The As host in coal is dominantly pyrite but also organic As can be of importance (Yudovich and Ketris 2005). Wendland et al. (1999) determined a mean As concentration of $55 \ \mu g \ g^{-1}$ in Lower Permian ("Rotliegend") basin sediments in Saxony (area 8) and correlated elevated values to groundwater As anomalies. In Miocene lignite from northeastern Bavaria (area 9), Dill and Wehner (1999) detected

163 89 μg^{-1} As on average (Clarke value for lignite: $7.4 \pm 1.4 \, \mu g \, g^{-1}$, Yudovich and Ketris 2005). In sin-164 gle wells in the Spessart (northwestern Bavaria, area 10), a correlation between elevated groundwater 165 As and the distribution of Upper Permian ("Zechstein") sediments was observed by Scheid et al. 166 (1999). Lower Triassic ("Buntsandstein") sandstones containing plant fossils in the Trier Basin (area 167 11) were shown to partly have high As concentrations of up to 880 µg g⁻¹ (Wagner 1999). Heinrichs and Udluft (1999) found that the distribution of extensively elevated As concentrations in Middle 168 169 Franconian (northern Bavaria, area 13) groundwater (up to 150 µg L⁻¹) is dependent on depositional 170 aquifer facies: only terrestrial sediments of the Upper Triassic ("Keuper") seem to produce groundwa-171 ter As in excess of the drinking water standard. In southern Bavaria, Middle Miocene terrestrial sands 172 ("Obere Süßwassermolasse", area 14) contain Fe hydroxide concretions with up to 1900 µg g⁻¹ As that can locally lead to elevated groundwater concentrations when reducing conditions occur (Bayer and 173 174 Henken-Mellies 1998). Accumulation processes of As in Middle Jurassic ("Dogger") sedimentary Fe 175 ores in the Upper Rhine Graben/Baden-Wuerttemberg (area 16) were described by Banning and Rüde 176 (2010): shallow marine environments fostering ooidic Fe ore formation provide conditions for synge-177 netic As accumulation. The studied depositional conditions proved suitable for As enrichment in 178 mainly goethitic Fe ooids during condensed sedimentation. Thereby, As accumulation is preferred 179 over other trace elements. 180

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Studies on sedimentary aquifer U and associated concentrations in groundwater derived from waterrock-interaction in Germany were scarce until about ten years ago. In contrary, the environmental impact of former extensive U mining, especially in Eastern Germany, has been characterized in detail (e.g. Wolkersdorfer 1996, Winkelmann et al. 2001, Baborowski and Bozau 2006). More recently, the inventory of geogenic background U in groundwater attracted more attention due to the new drinking water limitation. Birke et al. (2010) found a median of 0.17 µg L⁻¹ U in 908 German bottled water samples (maximum: 16 µg L-1). They statistically found that Triassic sandstone and crystalline basement aquifers (mainly Black Forest) represent the main hosts for elevated groundwater U. Hessian environmental authorities detected concentrations above 10 µg L⁻¹ U in 2.7% of 965 analyzed wells with a maximum concentration of 86 µg L⁻¹, elevated values were ascribed to geogenic input from Triassic rocks or Holocene peat deposits (HLUG 2008). For the SW German federal state of Baden-Württemberg, Liesch et al. (2015) found groundwater U concentrations controlled by geology with highest mean values associated to Upper Triassic aquifers. Uraniferous Upper Triassic sediments are also known from northern Bayaria and refered to as "active arkoses" (Abele et al. 1962). These syndiagenetic carbonate fluorapatite inclusions in the Keuper aquifer sandstones contain up to 1070 µg g⁻¹ U and were found to show structurally (CO₃²⁻ substitution in the crystal structure) and radiatively (αrecoil damage from U decay) enhanced mineral solubility. Extraction experiments indicated U release to groundwater during weathering: apatite alteration was identified as the responsible mechanism for widespread groundwater U concentrations in the region (16% of wells >10 μg L⁻¹) (area 13, Banning and Rüde 2015). Further north, the Keuper aquifer system hosts dolomitic inclusions instead of phosphates. Mineralogical and extraction data revealed that also this facies has a significant potential to release U to groundwater (area 12, Steffanowski and Banning 2017); the two uraniferous facies control the geogenic U groundwater problem in northern Bavaria. The subsurface of the "Mechernich Triassic triangle" in the far west of Germany consists of Buntsandstein sediments. Groundwater in this Lower Triassic aguifer contains up to 56 µg L⁻¹ U, with 7% of all samples in the study dataset exceeding the German guideline value (area 6, Schindler et al. 2016). Further studies describe anthropogenic mobilisation of geogenic subsurface U pools by agricultural activity in Mecklenburg-West Pomerania/NE Germany (van Berk and Fu 2016) and southern Bavaria (area 15, Banning et al. 2013b), drinking water treatment in North Rhine-Westphalia (Banning et al. 2017) or managed aquifer recharge on the island of Heligoland (area 1, Meurer and Banning 2018). The latter is basically a Buntsandstein rock in the North Sea, uplifted by salt tectonics. In the otherwise widely homogeneous red sandstones, Cu- and U-bearing cavity fillings and "fish eyes" - round reduction spots, partly with a black U-rich centre - can be found. They represent the source of geogenic U mainly bound in the carbonate fraction. On Heligoland, brackish water with partly elevated U concentrations is treated via reverse osmosis for use as drinking water. Rain water is irrigated to recharge the aquifer, an artificial accumulation of Buntsandstein debris. The fresh water dissolves part of the geogenic U pool in the sandstones, leading to U groundwater peaks during low tide.

This work revisits and extends nine earlier regional studies which unravelled geogenic As and/or U accumulation processes in aquifers with different geological and stratigraphical background, and their timings in geological history. These aquifers were hydrogeochemically, mineralogically and genetically characterized, they are actually or potentially affected by naturally elevated groundwater concentrations of As and/or U. Besides the major importance of the obtained results on a regional scale and global transferability based on comparable conditions between single areas, more may be learned from geochemical comparison of all study areas and their distribution in space and time. Therefore, the impact of sediment provenance and the spatio-temporal development of Europe on geogenic As and U distribution are evaluated in this study. Geochemical proxies derived from so far unpublished trace element data will be assessed to characterize the studied sediments' provenance, combine the results with the earlier As/U-focused studies, and derive a geodynamic explanation for large-scale natural As and U distribution in Germany.

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2 Materials and Methods

- A total of 270 sediment samples from nine study areas (eight studies in Germany, one in Mexico, Fig.
- 232 1) were considered in the present study, an overview is given in Table 1. Samples were taken from
- outcrops and boreholes, details about sediment sampling, geological and hydrogeological conditions,
- as well as regional As and/or U occurrence can be found in the papers cited in Figs. 1 and 2.

Table 1: overview of investigation areas and samples used in this study (in stratigraphical order, cf. Fig. 1 for location of investigation areas).

area no.	n=	stratigraphy	main lithology; As/U host phases
15	46	Neogene/Quaternary	Miocene alluvial sands, Quaternary fluvial gravels; Lignite,
			Fe hydroxides
5	36	Paleogene/Neogene	little consolidated marine sands, lower reduced and upper
			oxidized facies; pyrite, Fe hydroxide concretions
17	30	Paleogene/Neogene	terrestrial volcano-sedimentary basin filling (little consoli-
			dated sands, playa lake sediments); carbonates, Fe
			hydroxides
2	38	Upper Cretaceous	little consolidated marine sands, lower reduced and upper
			oxidized facies; pyrite, Fe hydroxide concretions, siderite
			concretions
16	32	Middle Jurassic	shallow marine limestone, mudstone, sandstone, ooidic Fe
			ores; Fe (hydr)oxides
12	14	Upper Triassic	terrestrial sandstone; uraniferous dolomite inclusions
13	47	Upper Triassic	terrestrial sandstone; uraniferous apatite inclusions, Fe
			oxides
1	18	Lower Triassic	red terrestrial sandstone; carbonate cavity fillings, U-rich
			black reduction spots
6	9	Lower Triassic	red terrestrial sandstone; Fe hydroxides

All samples were ground to powder grain size and analysed for bulk rock trace element geochemistry using Instrumental Neutron Activation Analysis (INAA). A sample aliquot of 1 g was encapsulated in a polyethylene vial and irradiated along with flux wires at a thermal neutron flux of 7*1012 n cm⁻² s⁻¹. ²⁴Na was allowed to decay for 7 days. Subsequently, the samples were counted on a high purity Ge detector with resolution of better than 1.7 KeV for the 1332 keV ⁶⁰Co photopeak. These analyses were performed by ISO 17025:2017-accredited Activation Laboratories Ltd., Ancaster, Ontario/Canada. For the trace element data presented here, analytical detection limits were 0.5 µg g⁻¹ for La, U and As; 0.2 µg g⁻¹ for Th and 0.1 µg g⁻¹ for Sc. For quality control, flux wires and control standards were used to compare decay-corrected activities to a calibration developed from certified international reference materials: DMMAS 108, 108-B, 111 or 119. One standard was run for every 11 samples. Replicates and blanks were analysed to check accuracy and precision of the data.

3 Results and Discussion

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3.1 Regional significance and comparative overview

Important outcomes from five of the nine regional studies evaluated in the course of the present work are visualized in Fig. 2 whereby As and U sources and sinks structures including mobilization and immobilization processes in the studied areas are assessed. Studies 13, 16, 2 and 5 mentioned in Fig. 2 were conducted in different parts of Germany (Fig. 1, chapter 1.3). Study 17 (Banning et al. 2012) discovered elevated groundwater As and U concentrations in the Villa de Reyes Graben, San Luis Potosí/central Mexico and succeeded in unravelling their common occurrence as well as the evolution of both elements' geochemical signatures. Absolute concentrations, inter-elemental behaviour and geochemical proxies supported the aforementioned view that both U and As are incompatible elements in magmatic differentiation and therefore accumulated in felsic lithologies. Dissolution of volcanic glass was identified as the main common release mechanism of U and As in the volcano-sedimentary Mexican basin. It represents input from a common source into groundwater, supported by geochemical signatures (e.g., normalized REE patterns) in volcanic rocks, basin fill sediments and groundwater (Banning et al. 2012). While As and U can both occur in groundwater of sedimentary basins, it became obvious that in most samples, only one of the trace elements showed high concentrations. This is explainable by separation due to redox heterogeneity in the aquifer, or by additional preferential input from secondary sedimentary sources (in this case preferential As over U input due to desorption from Fe (hydr)oxides and decarbonatization from caliche deposits).

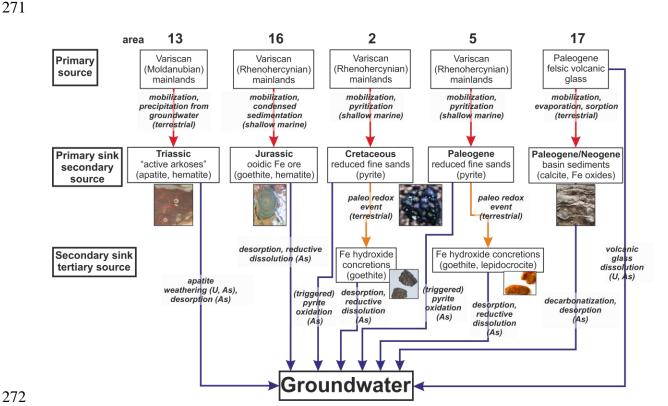


Fig. 2: Graphical overview of As and U sources and sinks structures including mobilization and immobilization processes in some of the studied areas. Area 13: Banning and Rüde (2015), area 16: Banning and Rüde (2010), area 2: Banning et al. (2013), area 5: Banning and Rüde (2010), area 17: Banning et al. (2012).

Figure 2 indicates that trace element accumulation processes and reservoirs can be of different orders such that spatio-temporal structure models of sources and sinks, and transfer processes between them, can be developed for a given area. The choice of study areas representing a variety of geological backgrounds and evolution make approaches and obtained results transferable to other regions. Sandstone-hosted uraniferous phosphates (area 13), shallow marine ooidic Fe ores (area 16), reduced glauconitic/pyritic marine sandy sediments (areas 2 and 5) and felsic volcano-sedimentary basins (area 17) are widespread geological environments worldwide and often share genetic and geochemical characteristics. This is especially supported by striking similarities between the studies in Cretaceous and Paleogene sediments (areas 2 and 5) exhibiting close analogies not only in geological situation and development but also in As distribution and behaviour in spite of a time span of ~50 Ma between sediment deposition in both areas.

3.2 Geodynamic model for large-scale natural As and U distribution in Germany

Besides the major importance of the obtained results on a regional scale and global transferability based on comparable conditions between single areas, more may be learned from general geochemical comparison of all study areas and their distribution in space and time. Therefore, the impact of sediment provenance and the spatio-temporal development of Europe on geogenic As and U distribution are evaluated in the following. Geochemical proxies are assessed to characterize the studied sediments' provenance and combine the results with the As and U accumulation processes derived in earlier studies. Figure 3 illustrates a ternary plot of La-Th-Sc to deduce source rock geochemistry and tectonic setting of the studied sediments.

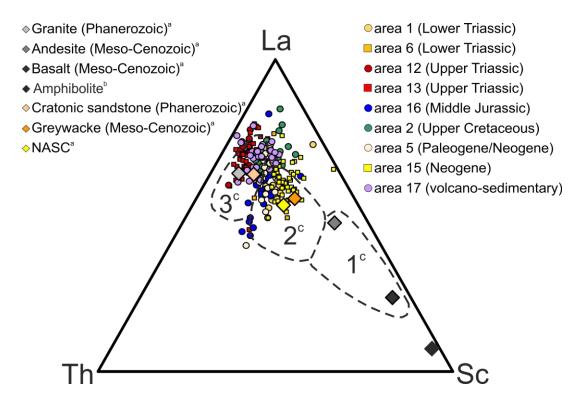


Fig. 3: La-Th-Sc plot of all studied sediment samples (cf. Fig. 1 for study area allocation). For comparison, plots of typical lithologies were implemented after aCondie (1993) and bCullers (1994). Furthermore, fields in the diagram after Bhatia and Crook (1986) indicate provenance rock tectonic setting with 1: oceanic island arc, 2: continental island arc, 3: active or passive continental margin. NASC: North American Shale Composite.

Most studied sediments plot in a rather dense cloud near the La corner of the diagram. Regarding the distal position of the samples from standard amphibolite, basalt and andesite (characterized by higher contents of compatible Sc) and the proximity to granite (higher degree of incompatible La and Th), a generally rather felsic provenance is obvious for all sediments (Fig. 3). Also in comparison to NASC, the North American Shale Composite (representative of upper crustal sediments; e.g., Gromet et al. 1984), the majority of samples plot in the more felsic range and are in this respect rather comparable to typical cratonic sandstone plotting within the data aggregation.

Concerning tectonic setting of the provenance sensu Bhatia and Crook (1986), most samples indicate derivation from continental margins with an overlap to the field of continental island arcs (Fig. 3), thus reflecting the tectonic configuration of Variscan central Europe as will be discussed later on. Sediment origin from oceanic island arcs can be excluded which was expectable regarding the choice of study areas.

To better differentiate between the single study areas with regard to geochemical provenance, a binary plot of incompatible La vs. compatible Sc is presented in Fig. 4.

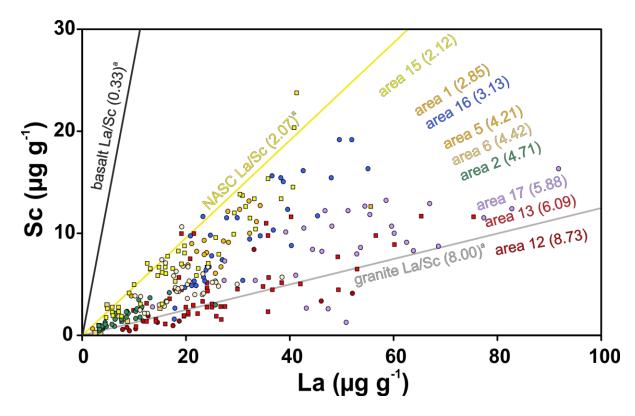


Fig. 4: La vs. Sc scatter plot and average La/Sc ratios for sediment samples from the nine study areas, cf. Fig. 1 for study area allocation. Basalt, NASC and granite standard ratios after ^aCondie (1993).

The observation of geochemical proxies indicating rather felsic provenance, also in comparison to average upper crustal sediment (NASC), is supported by the binary plot (Fig. 4). In spite of partly considerable scattering (R² of the nine data subsets: 0.11-0.72), differences between the single study areas become evident. Upper Triassic sediments from areas 12 (La/Sc ratio derived from trendline equation: 8.73) and 13 (6.09) as well as volcano-sedimentary basin filling from San Luis Potosí (area 17, 5.88) exhibit clearly more felsic provenance than marine sediments of Jurassic (area 16), Paleogene (area 5), Cretaceous (area 2) and Lower Triassic (areas 1 and 6) age. Neogene sediments from the southern German Molasse basin (area 15) document average La/Sc ratios very closely to the NASC signature.

335 signature

Consequently, areas with high U potential in terms of accumulation in rocks/sediments and elevated concentrations in groundwater originate from more felsic sources than sediments hosting "only" As.

The following is hypothesized:

a) Sedimentary environments potentially fostering As enrichment require above-average (i.e., above NASC) felsic provenance to assure sufficient supply of the incompatible trace element.

b) Sedimentary environments potentially fostering U enrichment require highly felsic provenance to assure sufficient supply of the (more) incompatible trace element, i.e. highly felsic origin is a prerequisite for sedimentary systems to create both, high As *and* U. Of course, con-

ditions of redeposition, climate, hydrochemistry and microbiology (let alone anthropogenic activity) control potential final groundwater concentrations.

c) Classical geochemical proxies like those presented here may serve as pre-diagnostic tools to characterize geochemical provenance (and thus, primary trace element sources) of sedimentary areas potentially affected by As and/or U.

In an effort to combine this approach and the previously discussed As/U accumulation processes, plots of La vs. U and La vs. As are presented as Fig. 5 and Fig. 6, respectively.



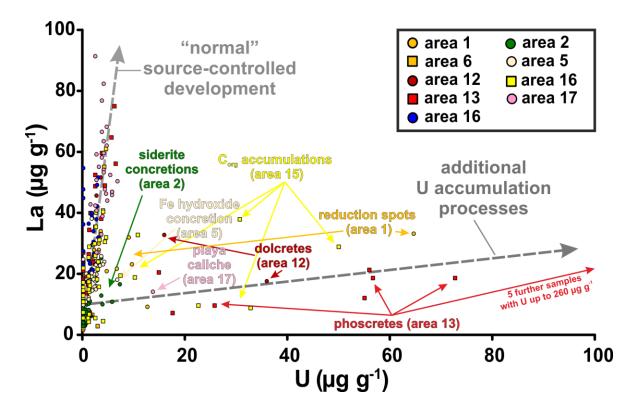


Fig. 5: U vs. La scatter plot for all studied sediments (cf. Fig. 1 for study area allocation), and secondary U accumulation mechanisms in sediment basins (cf. Fig. 2 and chapter 1.3).

The majority of samples in Fig. 5 exhibit a development along a positively trending line interpreted to represent "normal" source (i.e. provenance)-controlled behaviour. Here, common successive increase of both incompatible trace elements is obvious. Offsets from this "background" development, i.e. excess U, reflect additional sedimentary enrichment processes in the different areas. Slight accumulation was found in siderite concretions within Cretaceous fine sands (area 2). Uranium sorption to the purest Fe hydroxide concretion in oxidized sands of the Lower Rhine Embayment (area 5, this sample has the singular highest As content (1860 µg g⁻¹) of all study area) was not able to accumulate more than 7.2 µg g⁻¹ U. This finding supports the hypothesis that a more felsic provenance is needed for sufficient U supply – U availability was just too low in the Paleogene sediments such that not even this concretionary best option for accumulation could be taken advantage of. Evaporative concentration accounts for

elevated U contents in relic playa lake caliches (area 17). Apatite precipitation from U-bearing groundwater on former playa carbonates resulted in high U contents in phoscretes (area 13), also the more proximal original dolcretes (area 12) show some U accumulation. Very localized reducing conditions lead to punctual U hot spots in Lower Triassic sandstone on Heligoland (area 1), this phenomenon is also known from e.g., Switzerland (Hofmann 1990, Burkhalter 1995). Uranium accumulation in Neogene sediments (area 15) is almost exclusively observed in postsedimentary, i.e., Holocene, C_{org}-rich deposits like peat which successively concentrated U from groundwater.

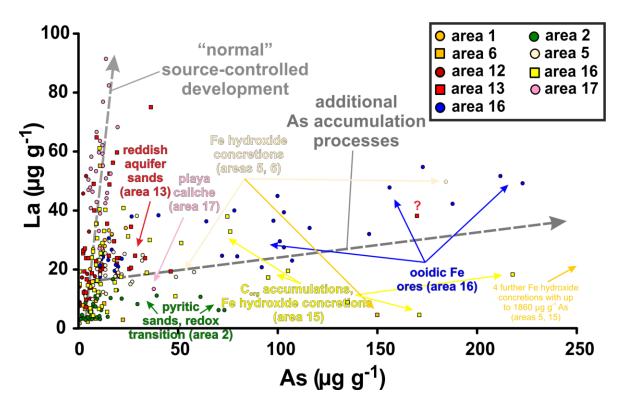


Fig. 6: As vs. La scatter plot for all studied sediments (cf. Fig. 1 for study area allocation), and secondary U accumulation mechanisms in sediment basins (cf. Fig. 2 and chapter 1.3).

Similar observations can be made in the As vs. La plot (Fig. 6) – a provenance-driven "background" development is overlain by different additional As concentration processes occurring in the studied sediment basins: sorption to Fe hydroxide coatings of silicate aquifer material (areas 6 and 13), accumulation in ooidic Fe ores during condensed sedimentation (area 16), fixation in pyrite and subsequent redistribution to redox transition zones and Fe hydroxide concretions (areas 2 and 5), enrichment by evaporation in playa caliches (area 17), concentration by sorption to organic material and in Fe(III) concretions (area 15). Jurassic ooidic Fe ores (area 16) with high As values only yield U contents of 1.5 µg g⁻¹ on average although depositional environment and potential host phases seem suitable for higher accumulation. Again, U availability in the sediment provenance is suggested to be the limiting factor for this phenomenon. Compared to the U-La plot (Fig. 5), the trend of "normal" source-controlled development between As and La for all samples appears less distinct. Obviously, the signa-

ture of the more incompatible U regarding geochemical provenance proxies is clearer. Moreover, additional redistribution processes after sediment deposition (like paleo redox events) seem to preferentially fractionate As, thereby altering primary geochemical signals.

To pursue these observations and possibly further support the formulated hypotheses, the stratigraphical and spatial distribution of the study areas will be assessed in the following and set in relation to their sediment provenance areas. Figure 7 offers an overview of the study areas' stratigraphical situation and the plate tectonic framework.

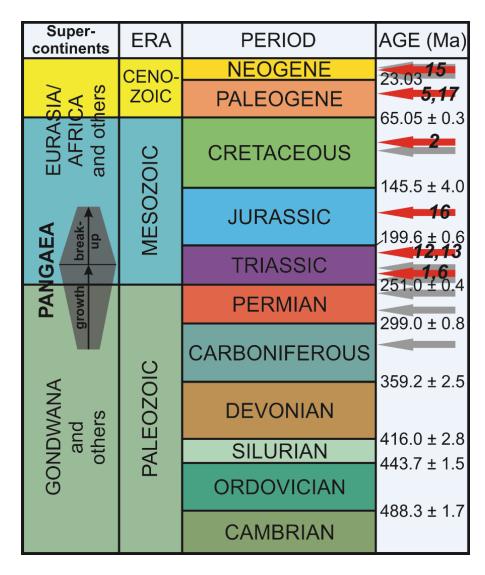


Fig. 7: Stratigraphical distribution of study areas (red arrows) and studied occurrences of sedimentary As in Germany from literature (grey arrows, cf. Fig. 1), absolute stratigraphical boundary ages (right side) after ICS (2009). Additionally, major crustal development (supercontinents, left side) after Bahlburg and Breitkreuz (1998) is indicated.

 $\begin{array}{c} 405 \\ 406 \end{array}$

The stratigraphical distribution of studied elevated sedimentary As in Germany reflects affected units ranging from Carboniferous to Neogene in age. A clustering of cases in Permo-Triassic is conspicuous which includes Germany's most significant As (and U) problem area in Northern Bavaria (areas 12,

410 13). Sediments from these periods often represent cratonic sandstones (cf. Fig. 3) derived from weath-411 ering of the then young Variscan massifs. Surveys of groundwater U conducted so far identified the 412 Triassic as "hot spot" age for affected aquifers (Birke et al. 2010, Kurth 2010, HLUG 2008, LGL and 413 LfU 2007; cf. capter 1.3). Likewise, an overview of As in German groundwater (Heinrichs and Udluft 414 1996) reports elevated concentrations mainly in Permo-Triassic aquifers. During that time, the for-415 mation climax of the supercontinent Pangaea as a result of the collision between Laurasia and 416 Gondwana was reached (Bahlburg and Breitkreuz 1998, Fig. 7). This process was accompanied by the 417 Variscan Orogeny representing a mountain-building event, relics of which are found today as e.g., the 418 Rhenish Massif, the Black Forest or the Bohemian Massif in Germany and neighbouring countries. 419 Denudation of the orogen and transport of resulting material to sediment basins started in the Upper 420 Carboniferous. Concluding from Fig. 7, only stratigraphies younger than that appear to be affected by 421 elevated As. Moreover, the upper limit of a stratigraphical "window" of sedimentary As occurrences 422 seems to be defined by the end of the Neogene period. Possible explanations for that must be sought in 423 the spatio-temporal geological development of Germany. 424 Figure 1 illustrates the spatial distribution of studied sedimentary As as well as Variscan terrains and 425 thick Quaternary overburden. It becomes evident that all known cases of elevated sedimentary As in 426 Germany are located south of the thick unconsolidated Quaternary cover governing geology and land-427 scape of northern Germany (except for the Triassic rock island Heligoland, area 1, which is not 428 covered by Quaternary). These mostly Pleistocene glacial and periglacial deposits with main prove-429 nance areas in northern and northeastern Europe (Henningsen and Katzung 2002) obviously did not 430 bring a significant As (or U) potential along. Not only would these observations explain the northern 431 boundary of As study areas (Fig. 1) but also raise expectations concerning better shallow groundwater 432 quality with respect to As and U in northern Germany. Indeed, Kunkel et al. (2004) characterized nat-433 ural groundwater quality in Germany and found in an evaluation of a large As dataset (n=1661) from 434 sand and gravel aquifers of the North German Plain (widely identical with the area north of the red line indicated in Fig. 1) 90th percentile values of 3.3-3.5 µg L⁻¹ As with hardly any single values above 435 436 10 µg L⁻¹ in different aquifer depth intervals of up to 50 m below ground surface. Supporting this in 437 their earlier survey, Heinrichs and Udluft (1996) do not report on geogenic groundwater concentra-438 tions in exceedance of 10 µg L⁻¹ As from this area. 439 The organisation foodwatch collected and published drinking water U data from German federal state 440 authorities (foodwatch 2009). In the dataset, they found that concentrations in the northern and central 441 federal states Brandenburg, Berlin, Bremen, Hamburg, Schleswig-Holstein, Lower Saxony, North 442 Rhine-Westphalia and Saarland are quantitatively below 10 µg L⁻¹ U (2 out of 427 samples in Meck-443 lenburg-Western Pomerania were above that limit) while southern federal states generally exhibit 444 more abundant violations of the drinking water guideline.

Another observation from Fig. 1 is the close spatial relation of most conducted sedimentary As and U studies to Variscan terrains in Germany supporting a relation between Variscan geology and trace

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element distribution as was already suspected from stratigraphy (Fig. 7). Differences between sedimentary occurrence of As and U in the single studied areas in this work can be derived from the distribution of Variscan granites and support the hypotheses on degree of provenance felsicity formulated earlier: Upper Triassic sediments from northern Bavaria (areas 12, 13) derive from weathering of the Bohemian Massif which belongs to the Moldanubian section of the European Variscides (like main parts of Black Forest and Vosges) representing the most metamorphic part with abundant felsic intrusions (Krebs 1976, Bahlburg and Breitkreuz 1998) while Cretaceous and Paleogene sediments (areas 2, 5) originate from the Rhenohercynian Variscan section with low degree of metamorphism and widely absent granitic intrusions (Fig. 1). At first glance, it seems surprising that the Jurassic sediments (area 16), being located in direct vicinity to the Moldanubian granite-rich Black Forest and Vosges (Fig. 1) do neither show highly felsic provenance (Fig. 4) nor elevated U potential (Fig. 5). Nevertheless, the Moldanubian crystalline terrains did not serve as sediment provenance but were submerged during the Middle Jurassic (Ziegler 1990), sediment (and Fe/As, but no U) input was accomplished from the Rhenish Massif (Sauer and Simon 1975) and thus, from the Rhenohercynian – a provenance of lower felsicity. To crosscheck this hypothesis, a short review of another basin that actually received its sedimentary filling from weathering of the Black Forest was conducted. Within the range of this Moldanubian terrain, only few rather small sedimentary basins filled with Upper Carboniferous to Permian material fulfill this condition, the biggest of which is the one around Baden-Baden in the northern Black Forest (Henningsen and Katzung 2002). Stefanian (Upper Carboniferous) partly Corg-rich arkoses and sandstones in that area, underlain by granite and derived from erosion of the proximal granitic young Variscan mountains, host several sedimentary U anomalies prospected in the past, the most significant of which is the deposit Müllenbach (Kneuper et al. 1977). The mineralization is of the roll-front type, main U carriers are uraninite and coffinite, bulk U₃O₈ contents of several wt.% are known. Arsenic is present in concentrations of up to 3000 µg g⁻¹, mainly as arsenopyrite. Kneuper et al. (1977) also found a positive correlation between As and U in the sediments. The authors genetically suggest common U and As precipitation from groundwater with trace elements primarily mobilized from Black Forest granites. This development is thus very similar to the one described here for Triassic phoscretes (area 13). Also Permian sediments in that basin partly show common enrichment of both U and As (Plinninger and Thuro 1999). Hydrothermal U mineralizations within granites are known from the southern Black Forest, e.g., near Menzenschwand. This site is the type locality of several secondary U-As minerals, one of which - Nielsbohrite [(KUO₂)₃(AsO₄)(OH)₄*H₂O] - was described by Walenta et al. (2009). These studies clearly underline Black Forest granites' high source potential for both U and As. The receiving sediment basins in the area are probably too small (in comparison to e.g., the South German Keuper Basin) to account for significant postdepositional redistribution processes and prominent groundwater As and U problems. Nevertheless, elevated U concentrations in

Black Forest mineral waters are known (Birke et al. 2010). Moreover, the thermal springs of Baden-

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Baden, recharged by meteoric waters circulating in the regional granites, show high As concentrations of 200 µg L⁻¹ on average; U presence is documented by U-bearing amorphous thermal Mn hydroxides and opaline sinters (Rüde 1996). In consequence, this short overview of a second Moldanubian massif fully supports the model of Variscan control over geogenic As and U distribution in Germany developed in the present work.

One exception among the existing studies of sedimentary As in Germany probably without Variscan influence should be mentioned: Bayer (1997) states that partly high As concentrations (up to 1900 µg⁻¹) in Fe-rich sections of the Bayarian molasse basin, esp. Badenean sands (area 14 in Fig. 1), originate, like the sediments themselves, from weathering of the Alpine Mountains; As is believed to primarily derive from mineralized zones in the eastern Alps. Uranium concentrations are low in the Fe concretions, Bayer (1997) detected <1 µg g-1 U in all his samples. According to the genetic model developed in the present work, this data argues for a primary trace element source of moderately high felsicity not able to supply sufficient U for enrichment in sediments. This source indeed is likely to be found in the eastern Alps which, in contrast to the central and western parts of the orogen, do not host significant granitic intrusions (Gwinner 1971). Nevertheless, several occasions of elevated U are known in groundwater and sediments from the range of the Miocene molasse basin (e.g. LGL and LfU 2007). An explanation for that can be given by another paleogeographical observation: after deposition of the Badenean sands, sediment provenance changed for the eastern and northern parts of the basin. During the subsequent Sarmatian and Pannonian, clastic material and groundwater supply were partly accomplished from the Bohemian Massif (Unger 1989) - a provenance of high felsicity with proven ability to produce U enrichment in associated sediment basins.

The conclusions drawn above may partly be transferable to other countries and even orogenies. In a survey of U occurrences in British groundwater, Smedley et al. (2006) found highest concentrations in sandstones derived from weathering of the Caledonian orogen (e.g. Old Red Sandstone) and speculate on U sorbed to abundant Fe (hydr)oxides as direct source for elevated groundwater U concentrations (no solid phases were studied). The primary source is likely to be the Caledonides which exhibit, akin to the German Moldanubian Variscides, numerous granite intrusions (e.g. Brown and Locke 1979).

Limitations of this study include the partly unequal distribution of the number of samples from the different study areas, ranging from 9 to 47, associated with heterogeneous levels of significance in statistical considerations. Also, trace element rich sample types (concretions, inclusions etc.) are likely to be overrepresented in the dataset and therefore do not resemble the composition of the whole aquifer/sediment system. Lastly, revisited studies were published within a decade (2010-2019), with partly differing people accomplishing sampling, sample preparation, sample analysis.

4 Conclusions

Geological evolution, expressed by geochemical proxies, can explain trace element distribution on different temporal and spatial scales, and help understand and forecast the occurrence of actual ans potential groundwater quality problem areas. Trace element abundance was shown to directly reflect supraregional and intra-basinal geological evolution. The distribution of areas with elevated As and U in Germany is large-scale widely determined by Variscan and Quaternary geology. Geochemical sediment provenance controls elevated As (felsic provenance) and U (highly felsic provenance) supply to sedimentary environments, whereby the different Variscan sections are decisive. Thus, magmatic geochemistry based on incompatibility of U and As is the ultimate control of trace element supply to sedimentary systems where subsequent intra-basinal processes of trace element accumulation, redistribution and eventually remobilization to groundwater take place and create the present-day situation.

The present work contributes to a deeper understanding of the interplay between geological history (magmatic and sedimentary), geochemistry, mineralogy, hydrogeology and hydrochemistry with respect to potentially hazardous geogenic trace elements, focussing on the still little understood situation in Germany. Increased comprehension of occurrence and behaviour of not only anthropogenic but also natural contaminants on a global scale will be necessary in the future in view of growing awareness regarding health impacts along with more stringent drinking water limitations, and continuously more severe population pressure on water resources, especially under climate change conditions.

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References

- Abele, G., Berger, K., Salger, M., 1962. Die Uranvorkommen im Burgsandstein Mittelfrankens [The uranium occurrences in the Burgsandstein of Middle Franconia]. Geologica Bavarica 19, 3-90 (in German).
- Acharyya, S.K., Lahiri, S., Raymahashay, B.C., Bhowmik, A., 2000. Arsenic toxicity of groundwater in parts of the Bengal basin in India and Bangladesh: the role of Quaternary stratigraphy and Holocene sea-level fluctuation. Environmental Geology 39, 1127-1137.

- 557 Armienta, M.A., Villaseñor, G., Rodriguez, R., Ongley, L.K., Mango, H., 2001. The role of arsenic-558 bearing rocks in groundwater pollution at Zimapán Valley, Mexico. Environmental Geology 559 40, 571-581.
- Armienta, M.A., Segovia, N., 2008. Arsenic and fluoride in the groundwater of Mexico. Environmental Geochemistry and Health 30, 345-353.
- Baborowski, M., Bozau, E., 2006. Impact of former mining activities on the uranium distribution in the River Saale (Germany). Applied Geochemistry 21, 1073-1082.
- Bahlburg, H., Breitkreuz, C., 1998. Grundlagen der Geologie [Basics of Geology]. Enke, Stuttgart, 328 pp (in German).
- Bajwa, B.S.; Kumar, S.; Singh, S.; Sahoo, S.K.; Tripathi, R.M. (2017): Uranium and other heavy toxic elements distribution in the drinking water samples of SW-Punjab, India. Journal of Radiation Research and Applied Sciences 10: 13-19.
- Banning, A., Coldewey, W.G., Göbel, P., 2009. A procedure to identify natural arsenic sources, applied in an affected area in North Rhine-Westphalia, Germany. Environmental Geology 57 (4), 775-787.
- Banning, A.; Rüde, T.R., 2010. Enrichment processes of arsenic in oxidic sedimentary rocks from geochemical and genetic characterization to potential mobility. Water Research 44, 5512-5531.
- Banning, A.; Cardona, A.; Rüde, T.R., 2012. Uranium and arsenic dynamics in volcano-sedimentary basins an exemplary study in north-central Mexico. Applied Geochemistry 27, 2160-2172.
- 577 Banning, A.; Rüde, T.R.; Dölling, B., 2013a. Crossing redox boundaries Aquifer redox history and effects on iron mineralogy and arsenic availability. Journal of Hazardous Materials 262, 905-914.
- Banning, A.; Demmel, T.; Rüde, T.R.; Wrobel, M., 2013b. Groundwater uranium origin and fate control in a river valley aquifer. Environmental Science & Technology 47, 13941-13948.
- Banning, A.; Rüde, T.R., 2015. Apatite weathering as a geological driver of high uranium concentrations in groundwater. Applied Geochemistry 59, 139-146.
- Banning, A.; Pawletko, N.; Röder, J.; Kübeck, C.; Wisotzky, F., 2017. Ex situ groundwater treatment triggering the mobilization of geogenic uranium from aquifer sediments. Science of the Total Environment 587-588, 371-380.
- Bayer, M., 1997. Natürliche Arsenanreicherungen in der Oberen Süßwassermolasse Bayerns [Natural arsenic accumulations in the Upper Süßwassermolasse of Bavaria]. Veröffentlichungen des Grundbauinstitutes der Landesgewerbeanstalt Bayern 77 (in German).
- Bayer, M., Henken-Mellies, W.-U., 1998. Untersuchung arsenführender Sedimente in der Oberen Süßwassermolasse Bayerns [Investigation of arsenic bearing sediments of the Upper Süßwassermolasse in Bavaria]. GeoCongress 4, 42-58 (in German).
- Berg, M., Tran, H.C., Nguyen, T.C., Pham, H.V., Schertenleib, R., Giger, W., 2001. Arsenic contamination of groundwater and drinking waterin Vietnam: a human health threat. Environmental Science and Technology 35 (13), 2621-2626.
- BfS (Bundesamt für Strahlenschutz, ed.), 2009. Strahlenexposition durch natürliche Radionuklide im Trinkwasser in der Bundesrepublik Deutschland [Radiation exposure by natural radionuclides in drinking water in the Federal Republic of Germany]. urn:nbn:de:0221-20100319945 (in German).

- Bhatia, M.R., Crook, K.A.W., 1986. Trace element characteristics of graywackes and tectonic setting discrimination of sedimentary basins. Contributions to Mineralogy and Petrology 92, 181-193.
- Birke, M., Rauch, U., Lorenz, H., Kringel, R., 2010. Distribution of uranium in German bottled and tap water. Journal of Geochemical Exploration 107 (3), 272-282.
- Bjørklund, G., Christophersen, O.A., Chirumbolo, S., Selinus, O., Aaseth, J. (2017). Recent aspects of uranium toxicology in medical geology. Environmental Research 156, 526-533.
- Bots, P., Behrends, T., 2008. Uranium mobility in subsurface aqueous systems: the influence of redox conditions. Mineralogical Magazine, 72 (1), 381-384.
- Brown, G.C., Locke, C.A., 1979. Space-time variations in British Caledonian granites some geophysical correlations. Earth and Planetary Science Letters 45, 69-79.
- Brown, C.J., Jurgens, B.C., Katz, B.G., Landon, M.K., Eberts, S.M., 2007. Arsenic and uranium in four aquifer settings: Occurrence, distribution, and mechanisms for transport to supply wells.

 NGWA Naturally Occurring Contaminants Conference: Arsenic, Radium, Radon, Uranium,
 March 22-23, 2007, Charleston/U.S.A.
- Bundschuh, J., Farias, B., Martin, R., Storniolo, A., Bhattacharya, P., Cortes, J., Bonorinom, G., Albouy, R., 2004. Groundwater arsenic in the Chaco-Pampean Plain, Argentina: case study from Robles country, Santiago del Estero Province. Applied Geochemistry 19 (2), 231-243.
- Burkhalter, R.M., 1995. Ooidal ironstones and ferruginous microbialites: origin and relation to sequence stratigraphy (Aalenian and Bajocian, Swiss Jura mountains). Sedimentology 42, 57-74.
- 619 Condie, K.C., 1993. Chemical composition and evolution of the upper continental crust: Contrasting results from surface samples and shales. Chemical Geology 104, 1-37.
- Cremer, N., Obermann, P., Wisotzky, F., 2003. Mobilization of Nickel, Cobalt and Arsenic in a Multi-aquifer Formation of the Lower Rhine Valley: Identification and Modeling of the Processes
 Controlling Metal Mobility. In: Schulz, H.D., Hadeler, A. (Eds.) Geochemical Processes in
 Soil and Groundwater Measurement Modelling Upscaling (Proceedings GeoProc2002),
 3-18, Wiley, Weinheim.
- 626 Cullers, R.L., 1994. The chemical signature of source rocks in size fractions of Holocene stream sedi-627 ment derived from metamorphic rocks in the Wet Mountains region, Colorado, U.S.A. 628 Chemical Geology 113, 327-343.
- 629 Cuney, M. (2010): Evolution of Uranium Fractionation Processes through Time: Driving the Secular Variation of Uranium Deposit Types. Economic Geology (2010) 105 (3): 553–569.
- Del Razo, L.M., Arellano, M.A., Cebrián, M.E., 1990. The Oxidation States of Arsenic in Well-Water from a Chronic Arsenicism Area of Northern Mexico. Environmental Pollution 64 (2), 143-153.
- Dill, H.G., Wehner, H., 1999. The depositional environment and mineralogical and chemical compositions of high ash brown coal resting on early Tertiary saprock (Schirnding Coal Basin, SE Germany). International Journal of Coal Geology 39, 301-328.
- Doi, K., Hirono, S., Sakamaki, Y., 1975. Uranium Mineralization by Groundwater in Sedimentary Rocks, Japan. Economic Geology 70, 628-646.
- Foodwatch e.V., 2009. Uran-Werte im Trinkwasser (Stand Herbst 2009) [Uranium concentrations in drinking water, status: autumn 2009]. http://foodwatch.de/kampagnen__themen/mineralwas ser/trinkwasser/index_ger.html (in German).

- Frengstad, B., Skrede, A.K.M., Banks, D., Krog, J.R., Siewers, U., 2000. The chemistry of Norwegian groundwaters: III. The distribution of trace elements in 476 crystalline bedrock groundwaters, as analysed by ICP-MS techniques. The Science of the Total Environment 246, 21-40.
- 645 Giblin, A.M., Batts, B.D., Swaine, D.J., 1981. Laboratory simulation studies of uranium mobility in natural waters. Geochimica et Cosmochimica Acta 45, 699-709.
- Goldberg, G., Lepper, J., Röhling, H.G., 1995. Geogene Arsengehalte in Gesteinen und Grundwässern des Buntsandsteins in Südniedersachsen [Geogenic arsenic contents in rocks and groundwater of the Buntsandstein formation in southern Lower Saxony, Germany]. Zeitschrift für Angewandte Geologie 41, 118-124 (in German).
- 651 Gromet, L.P., Dymek, R.F., Haskin, L.A., Korotev, R.L., 1984. The "North American Shale Compo-652 site": Its compilation, major and trace element characteristics. Geochimica et Cosmochimica 653 Acta 48, 2469-2482.
- 654 Gwinner, M.P., 1971. Geologie der Alpen Stratigraphie, Paläogeographie, Tektonik [Geology of the Alps stratigraphy, paleogeography, tectonics]. Schweizerbart, Stuttgart, 477 pp (in German).
- Heinrichs, G., Udluft, P., 1996. Geogenes Arsen in Grundwässern Deutschlands unter Berücksichtigung der Aquifergeologie [Geogenic arsenic in German groundwater with respect to aquifer geology]. Zeitschrift der Deutschen Geologischen Gesellschaft 147, 519-530 (in German).
- Heinrichs, G., Udluft, P., 1999. Natural arsenic in Triassic rocks: A Source of drinking water contamination in Bavaria, Germany. Hydrogeology Journal 7, 468-476.
- Henningsen, D., Katzung, G., 2002. Einführung in die Geologie Deutschlands [Introduction to the geology of Germany]. Spektrum, Berlin (in German).
- Herch, A., 1997. Untersuchungen zur hydrogeochemischen Charakteristik der Spurenelemente und Schwefelspezies im Aachener Thermalwasser [Study on hydrogeochemical characteristics of trace elements and sulphur species in the thermal water of Aachen]. Mitteilungen zur Ingenieurgeologie und Hydrogeologie 64, RWTH Aachen University (in German).
- 668 HLUG (Hessisches Landesamt für Umwelt und Geologie, ed.), 2008. Urankonzentrationen in 669 Hessischen Grundwässern – eine Studie des Hessischen Landesamtes für Geologie und 670 Umwelt [Uranium concentrations in Hessian groundwater – a study of the Hessian authority 671 for Geology and Environment] (in German).
- Hofmann, B.A., 1990. Reduction spheroids from northern Switzerland Mineralogy, geochemistry and genetic models. Chem. Geol. 81, 55-81.
- 674 ICS (International Commission on Stratigraphy, ed.), 2009. International stratigraphic chart.
- Katsoyiannis, I.A., Hug, S.J., Ammann, A., Zikoudi, A., Hatziliontos, C., 2007. Arsenic speciation and uranium concentrations in drinking water supply wells in Northern Greece: Correlations with redox indicative parameters and implications for groundwater treatment. Science of the Total Environment 383 (1-3), 128-140.
- Kipp, G.G., Stone, J.J., Stetler, L.D., 2009. Arsenic and uranium transport in sediments near abandoned uranium mines in Harding County, South Dakota. Applied Geochemistry 24, 2246-2255.
- Kneuper, G., List, K.-A., Maus, H., 1977. Geologie und Genese der Uranmineralisation des Oostroges im Nordschwarzwald [Geology and genesis of the uranium mineralization of the Oostrog in the northern Black Forest]. Erzmetall 30 (11), 522-525 (in German).

- Krebs, W., 1976. The Tectonic Evolution of Variscan Meso-Europe, In: Ager, D.V., Brooks, M. (eds.), Europe from Crust to Core. Wiley, London.
- Kumar, A.; Karpe, R.K.; Rout, S.; Gautam, Y.P.; Mishra, M.K.; Ravi, P.M.; Tripathi, R.M. (2016):
 Activity ratios of ²³⁴U/²³⁸U and ²²⁶Ra/²²⁸Ra for transport mechanisms of elevated uranium in
- alluvial aquifers of groundwater in south-western (SW) Punjab, India. Journal of Environmental Radioactivity 151: 311-320.
- Kumar, A.; Singh, C.K. (2020): Arsenic enrichment in groundwater and associated health risk in Baridoab region of Indus basin, Punjab, India. Environmental Pollution 256, 1-13.
- Kunkel, R., Voigt, H.-J., Wendland, F., Hannappel, S., 2004. Die natürliche, ubiquitär überprägte Grundwasserbeschaffenheit in Deutschland [The natural, ubiquitarily overprinted, groundwater quality in Germany]. Schriften des Forschungszentrums Jülich, Reihe Umwelt/Environment 47, 204 pp, Jülich.
- Kurth, D., 2010. Distribution and origin of geogenic uranium in groundwater systems of the southern Lower Rhine Embayment and adjacent areas. B.Sc. thesis, RWTH Aachen University, Germany (unpublished).
- Lapworth, D.J.; Krishan, G.; MacDonald, A.M.; Rao, M.S. (2017): Groundwater quality in the alluvial aquifer system of northwest India: new evidence of the extent of anthropogenic and geogenic contamination. Science of the Total Environment 599-600: 1433-1444.
- LGL, LfU (Bayerisches Landesamt für Gesundheit und Lebensmittelsicherheit, Bayerisches Landesamt für Umwelt, Eds.), 2007. Untersuchungen zum Vorkommen von Uran im Grundund Trinkwasser in Bayern [Study of uranium occurrence in groundwater and drinking water in Bayaria] (in German).
- Lowers, H.A., Breit, G.N., Foster, A.L., Whitney, J., Yount, J., Uddin, M.N., Muneem, A.A., 2007.
 Arsenic incorporation into authigenic pyrite, Bengal Basin sediment, Bangladesh. Geochimica et Cosmochimica Acta 71, 2699-2717.
- Matschullat, J., 1999. Arsen in der Geosphäre [Arsenic in the geosphere]. In: Rosenberg, F., Röhling,
 H.G. (Eds.), Arsen in der Geosphäre. Schriftenreihe der Deutschen Geologischen Gesellschaft
 6, 5-20 (in German).
- Merkel, B., Sperling, B., 1998. Hydrogeochemische Stoffsysteme, Teil II [Hydrogeochemical substances, part II]. Schriftenreihe des Deutschen Verbandes für Wasserwirtschaft und Kulturbau (DVWK) 117, Bonn (in German).
- Mertens, J., 2000. Anreicherung und Bindungsformen geogener Arsen- und Schwermetall-belastungen glaukonitreicher Kreidesedimente im mittleren Ruhrgebiet [Enrichment and fractionation of geogenic arsenic and heavy metals in glauconite-rich Cretaceous sediments in the Middle Ruhr Area.]. M.Sc. thesis, University of Essen, Germany (in German).
- Meurer, M.; Banning, A., 2019. Uranmobilisierung im Helgoländer Buntsandstein Auswirkungen auf die Brack- und Trinkwasserqualität. Grundwasser 24, 43-50.
- Missana T., García-Gutiérrez, M., Maffiotte, C., 2003. Experimental and modelling study of the uranium(VI) sorption on goethite. Journal of Colloid and Interface Science 260, 291-301.
- Nicolli, H.B., Bundschuh, J., García, J.W., Falcón, C.M., Jean, J.-S., 2010. Sources and controls for the mobility of arsenic in oxidizing groundwaters from loess-type sediments in arid/semi-arid dry climates – Evidence from the Chaco-Pampean plain (Argentina). Water Research 44 (19),

727 5589-5604.

- Orloff, K.G., Mistry, K., Charp, P., Metcalf, S., Marino, R., Shelly, T., Melaro, E., Donohoe, A.M., Jones, R.L., 2004. Human exposure to uranium in groundwater. Environmental Research 94, 319-326.
- Oyarzun, R., Lillo, J., Higueras, P., Oyarzun, J., Maturana, H., 2004. Strong arsenic enrichment in sediments from the Elqui watershed, Northern Chile: industrial (gold mining at El Indio-Tambo district) vs. geologic processes. Journal of Geochemical Exploration 84, 53-64.
- Patnaik, R.; Lahiri, S.; Chahar, V.; Naskar, N.; Sharma, P.K.; Avhad, D.K.; Bassan, M.K.T.; Knolle, F.; Schnug, E.; Srivastava, A., 2015. Study of uranium mobilization from Himalayan Siwaliks to the Malwa region of Punjab state in India. Journal of Radioanalytical and Nuclear Chemistry 308: 913-918.
- Pedersen, H.D., Postma, D., Jakobsen, R., 2006. Release of arsenic associated with the reduction and transformation of iron oxides. Geochimica et Cosmochimica Acta 70, 4116-4129.
- Planer-Friedrich, B., Armienta, M.A., Merkel, B.J., 2001. Origin of arsenic in the groundwater of the Rioverde basin, Mexico. Environmental Geology 40, 1290-1298.
- Plinninger, R.J., Thuro, K., 1999. Die geologischen Verhältnisse beim Vortrieb des Meisterntunnels Bad Wildbad/Nordschwarzwald [Geological conditions during contruction of the Meisterntunnel Bad Wildbad/northern Black Forest]. Jahrbuch 1999 des Oberrheinischen Geologischen Vereins, 1-11 (in German).
- Ravenscroft P., Burgess W.G., Ahmed K., Burren M., Perrin J., 2005. Arsenic in groundwater of the Bengal Basin, Bangladesh: Distribution, field relations, and hydrogeological setting. Hydrogeology Journal 13, 727-751.
- Rosenberg, F., Mittelbach, G., Kirnbauer, T., 1999. Geogene Arsengehalte im Bereich der Wiesbadener Thermalquellen [Geogenic arsenic contents within the range of the thermal springs of Wiesbaden]. In: Rosenberg, F., Röhling, H.-G. (eds.): Arsen in der Geosphäre, Schriftenreihe der Deutschen Geologischen Gesellschaft 6, 101-106 (in German).
- Rüde, T.R., 1996. Beiträge zur Geochemie des Arsens [Contributions to arsenic geochemistry].
 Karlsruher Geochemische Hefte 10, 1-206 (in German).
- Sauer, K., Simon, P., 1975. Die Eisenerze des Aalenium und Bajocium im Oberrheingraben (Grube Kahlenberg, Grube Schönberg und kleinere Vorkommen) [Aalenian and Bajocian iron ores in the Upper Rhine Graben (Kahlenberg quarry, Schönberg quarry and smaller deposits)].

 Geologisches Jahrbuch D10, 25-68 (in German).
- Scheid, Y., Schiedek, T., Ebhardt, G., 1999. Geogenes Arsen in Quellwässern des Spessarts (Bayern)
 [Geogenic arsenic in spring waters of the Spessart (Bayaria)]. Schriftenreihe der Deutschen
 Geologischen Gesellschaft 6, 127-130 (in German).
- Schindler, F., Wisotzky, F., Banning, A., Cremer, N. (2016). Uranmobilisierung im Buntsandstein des
 Mechernicher Triasdreiecks [Uranium mobilization from Buntsandstein sediments of the
 Mechernic Triassic Triangle]. In: Blum, P., Goldscheider, N., Göppert, N., Kaufmann-Knoke,
 R., Klinger, J., Liesch, T., Stober, I. (eds.): Grundwasser Mensch Ökosysteme: 262-263.
 doi: 10.5445/KSP/1000051730 (in German).
- Sherman, H.M., Gierke, J.S., Anderson, C.P., 2007. Controls on spatial variability of uranium in sandstone aquifers. Ground Water Monitoring & Remediation 27 (2), 106-118.
- Smedley, P.L., Kinniburgh, D.G., 2002. A review of the source, behaviour and distribution of arsenic in natural waters. Applied Geochemistry 17, 517-568.

- Smedley, P.L., Smith, B., Abesser, C., Lapworth, D., 2006. Uranium occurrence and behaviour in British groundwater. British Geological Survey Commissioned Report, CR/06/050N.
- Spoljaric, N., Crawford, W.A., 1978. Glauconitic Greensand: A possible filter of heavy metal cations from polluted water. Environmental Geology, 2 (4), 215-221.
- Stanger, G., 2005. A palaeo-hydrogeological model for arsenic contamination in southern and southeast Asia. Environmental Geochemistry and Health 27, 359-367.
- Steffanowski, J.; Banning, A. (2017): Uraniferous dolomite a natural source of high groundwater
 uranium concentrations in northern Bavaria, Germany? Environmental Earth Sciences 76,
 508-518.
- Stewart, B.D., Mayes, M.A., Fendorf, S. (2010): Impact of Uranyl-Calcium-Carbonato Complexes on Uranium(VI) Adsorption to Synthetic and Natural Sediments. Environmental Science & Technology 44, 928-934.
- Stollenwerk, K.G., 2002. Geochemical Processes Controlling Transport of Arsenic in Groundwater: A Review of Adsorption. In: Welch, A.H. and Stollenwerk, K.G. (Eds.) Arsenic in Ground Water – Geochemistry and Occurrence, 67-100, Springer, New York.
- Taylor, S.R., 1964. Abundance of chemical elements in the continental crust: a new table. Geochimica et Cosmochimica Acta 28, 1273-1285.
- Unger, H.J., 1989. Die Lithozonen der Oberen Süßwassermolasse Südostbayerns und ihre vermutlich zeitlichen Äquivalente gegen Westen und Osten [Lithozones of the Upper Süßwassermolasse of southeastern Bavaria and its assumed equivalents towards the west and the east]. Geologica Bavarica 94, 195-223 (in German).
- Van Berk, W., Fu, Y. (2017). Redox Roll-Front Mobilization of Geogenic Uranium by Nitrate Input
 into Aquifers: Risks for Groundwater Resources. Environmental Science & Technology 51,
 337-345.
- Wagner, J.-F., 1999. Arsen im Grundwasser der triadischen Randfazies Luxemburgs [Arsenic in the triadic marginal facies of Luxembourg]. Schriftenreihe der Deutschen Geologischen Gesellschaft 6, 131-132 (in German).
- Walenta, K., Hatert, F., Theye, T., Lissner, F., Röller, K., 2009. Nielsbohrite, a new potassium uranyl arsenate from the uranium deposit of Menzenschwand, southern Black Forest, Germany. European Journal of Mineralogy 21, 515-520.
- Welch, A.H., Lico, M.S., 1998. Factors controlling As and U in shallow ground water, southern Carson Desert, Nevada. Applied Geochemistry 13 (4), 521-539.
- Wendland, A., Rank, G., Barth, A., 1999. Verteilung des Arsens in den Rotliegendsedimenten Sachsens [Arsenic distribution in Rotliegend sediments of Saxony]. Schriftenreihe der Deutschen Geologischen Gesellschaft 6, 93-100 (in German).
- Winkelmann, I., Thomas, M., Vogl, K., 2001. Aerial measurements on uranium ore mining, milling and processing areas in Germany. Journal of Environmental Radioactivity 53, 301-311.
- Wolkersdorfer, C., 1996. Hydrogeochemical investigations of an abandoned uranium mine in the Erzgebirge/Germany. Applied Geochemistry 11, 237-241.
- Yudovich, Y.E., Ketris, M.P., 2005. Arsenic in coal: a review. International Journal of Coal Geology 61, 141-196.
- Zahid, A., Hassan, M.Q., Breit, G.N., Balke, K.-D., Flegr, M., 2009. Accumulation of iron and arsenic in the Chandina alluvium of the lower delta plain, Southeastern Bangladesh. Environmental Geochemistry and Health 31, 69-84.

815	Zamora, M.L., Tracy, B.L., Zielinski, J.M., Meyerhof, D.P., Moss, M.A., 1998. Chronic Ingestion of
816	uranium in Drinking Water: A Study of Kidney Bioeffects in Humans. Toxicological Sciences
817	43, 68-77.
818	Ziegler, P.A., 1990. Geological Atlas of Western and Central Europe. Elsevier, Amsterdam.