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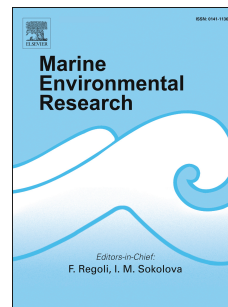
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# Accepted Manuscript

Neutral red retention time assay in determination of toxicity of nanoparticles

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1      **Neutral red retention time assay in determination of toxicity of nanoparticles**

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20     *Keywords: Mytilus, metal oxide, lysosome, membrane stability, neutral red, NRRT*

21

22     1

22

23 **Abstract**

24 The neutral red retention time (NRRT) assay is useful for detecting decreased  
25 lysosomal membrane stability in haemocytes sampled from bivalves, a phenomenon  
26 often associated with exposure to environmental pollutants including nanomaterials.  
27 Bivalves are popular sentinel species in ecotoxicology and use of NRRT in study of  
28 species in the genus *Mytilus* is widespread in environmental monitoring. The NRRT  
29 assay has been used as an *in vivo* test for toxicity of carbon nanoparticles (Moore MN,  
30 Readman JAJ, Readman JW, Lowe DM, Frickers PE, Beesley A. 2009. Lysosomal  
31 cytotoxicity of carbon nanoparticles in cells of the molluscan immune system: An *in*  
32 *vivo* study. *Nanotoxicology*. 3 (1), 40-45). We here report application of this assay  
33 adapted to a microtitre plate format to a panel of metal and metal oxide nanoparticles  
34 (2ppm). This showed that copper, chromium and cobalt nanoparticles are toxic by this  
35 criterion while gold and titanium nanoparticles are not. As the former three  
36 nanoparticles are often reported to be cytotoxic while the latter two are thought to be  
37 non-cytotoxic, these data support use of NRRT as a general *in vitro* assay in  
38 nanotoxicology.

39

## 40        **1. Introduction**

41            The unusual properties of nanomaterials provide them with several possible routes to  
42 toxicity in biological systems. Their small size sometimes enables them to cross important  
43 biobarriers e.g. skin, blood-brain, intestine, maternal-foetus (Tedesco and Sheehan, 2010;  
44 Elsaesser and Howard, 2012; Jiang et al., 2014). Their very large surface area to volume ratio  
45 enables a greater proportion of atoms to be displayed on the particle surface compared to  
46 corresponding macromaterials (Nel et al., 2009; Nel et al., 2013). Moreover, specific  
47 functional groups on nanoparticle surfaces may facilitate biospecific interactions allowing a  
48 range of possible biological effects (Hoet et al., 2004; Moore, 2006; Klaper et al., 2014).  
49 Nanomaterials can also translocate within the human body into other systems such as  
50 circulatory and lymphatic vessels (Gwinn and Vallyathan, 2006; Buzea et al., 2007; Elsaesser  
51 and Howard, 2012). Thus, nanoparticles have significant potential to cause adverse health  
52 effects in humans and other organisms upon prolonged exposure.

53            Because of increasing commercial production and use of nanomaterials, issues  
54 of their accumulation and fate in the environment and their possible effects on  
55 ecosystems arise (Moore, 2006; Tedesco and Sheehan, 2010; Ivask et al., 2014). The  
56 majority of human habitation worldwide is within 100km of coastlines and the aquatic  
57 environment collects domestic, agricultural, shipping and industrial runoffs from  
58 these coastal zones. This makes aquatic ecosystems particularly at risk to potential  
59 toxicity of nanomaterials of anthropogenic origin. Invertebrates are key elements of  
60 the aquatic food chain and mussels are amongst the most abundant of these (Baun et

61 al., 2008). As filter-feeders, mussels are exquisitely selective in the particle size-  
62 range which they ingest (Defosse and Hawkins, 1997; Ward and Kach, 2009) and  
63 can bioconcentrate metals and organic pollutants within their tissues. This has led to  
64 their widespread study in ecotoxicology (Moore, 1985; Widdows and Donkin, 1992)  
65 and filter-feeders have been suggested as especially attractive targets for probing the  
66 environmental fate of nanomaterials (Moore, 2006; Ward and Kach, 2009; Canesi et  
67 al., 2012).

68 Lysosomes are important subcellular organelles that contain many hydrolytic  
69 enzymes, carry out protein degradation and detoxify some foreign compounds. At the  
70 cellular level, lysosomal digestion pathways include phagocytosis, endocytosis and  
71 autophagy. The lysosomal membrane protects the cytosol, and therefore the rest of the  
72 cell, from leakage of degradative enzymes. However, malfunctioning of lysosomes  
73 and their accumulation of toxic pollutants have been linked to lysosomal storage  
74 diseases and result in lysosomal injury and oxidative damage, in some cases leading  
75 to cell death (Moore et al., 2007). The neutral red retention time (NRRT) assay takes  
76 advantage of this phenomenon by measuring decreased time of retention of a dye,  
77 neutral red (ACS no. 553-24-2), within phagocytic haemocytes of a range of aquatic  
78 organisms including mussels, crustaceans and fish (Regoli, 1992; Tedesco et al, 2008;  
79 Lowe et al 1995; Svendsen et al, 2004). In the popular sentinel species, *Mytilus edulis*,  
80 hemocytes are essential immune system components (Rickwood and Galloway,  
81 2004). NRTT has been reported as a useful indicator of the organism's overall health

82 status because animals exposed to pollutants often have compromised lysosomal  
83 stability (Moore et al., 2009; Borenfreund and Puerner 1985; Piola et al., 2013). A  
84 spectrophotometric version of the assay was developed by Babich and Borenfreund  
85 (1990) and a microscopic slide observation method was developed by Moore et al.,  
86 (2009). This assay takes advantage of the tendency of haemocytes to take up  
87 nanoparticles most probably by either phagocytosis or macro-endocytosis and  
88 involves exposing haemocytes to nanoparticles on a microscope slide (Moore et al.,  
89 2009). In this short report, we have adapted this methodology to a microtitre plate  
90 format enabling high-throughput screening of large numbers of replicates, doses and  
91 nanoparticles simultaneously (Fig. 1). As proof of principle, we have assessed a panel  
92 of metal and metal oxide nanoparticles with this assay.

93

## 94        **2. Materials and Methods**

### 95        *2.1. Mytilus edulis sampling*

96            *M. edulis* individuals (4-6cm shell-length) were collected from an intertidal site in  
97        Cork Harbour, Ireland (location: 51.49°N, 8 18°W; Lyons et al., 2003). All Animals  
98        were acclimated in tanks for a week with a 12 h light/dark cycle at a temperature of  
99        15°C and 34–36‰ salinity, fed and with regular changing of water.

100

### 101        *2.2. Nanoparticle suspension preparation*

102            Metal or metal oxide nanoparticles (copper oxide, titanium dioxide, gold,  
103        chromium oxide and cobalt oxide) of nominal sizes <50nm were purchased from  
104        Sigma-Aldrich (Dorset, UK). Nanopowders (10mg) were suspended in 10 ml of 20  
105        mM citric acid adjusted to pH 7, and sonicated for 1h using a tip sonicator. A stepped  
106        microtip was used and the total power transferred to the suspension was 2.4W  
107        (determined by the calorimetric method). Ultrasound was applied as 15s pulses with  
108        15s breaks between them (Taurozzi et al., 2010). The suspensions were left at 60°C  
109        overnight and were then filtered using a 220nm pore size cellulose acetate filter  
110        (Millipore, Watford UK).

111

### 112        *2.3. Exposure of haemolymph to nanoparticles*



113 Haemolymph samples were freshly extracted for NRRT assay as described by  
114 Moore et al. (2009). In the present work, haemolymph from each of five animals was  
115 extracted from adductor muscle using a 20 gauge hypodermic needle fitted on a 1 ml  
116 syringe containing 100µl tris buffered saline buffer, which was pooled to provide a  
117 total volume of 2 ml haemolymph solution. Three biologically independent replicates  
118 were used (i.e. haemolymph was taken from 3x5 individual animals). Samples were  
119 constantly vortexed to resuspend the haemolymph and prevent aggregation.  
120 Haemolymph was then evenly aliquoted (500 µL) followed by exposure to  
121 nanoparticles at a final concentration of 2 ppm for 1 h at ambient temperature (20°C).  
122 Tubes were gently shaken every 5 min to optimise exposure. The above procedure  
123 was applied to a panel of metal or metal oxide nanoparticles and a control sample was  
124 treated identically but without the presence of nanoparticle.

125

#### 126 *2.4. Neutral red retention time (NRRT) assay*

127 Following nanoparticle exposure, 100 µl haemolymph from all six treatment  
128 groups was loaded into individual wells of a 96-well microtitre plate (Sarstedt,  
129 Wexford Ireland). This was performed with three independent biological replicates.  
130 Fifty µl stock neutral red dye solution (200 µM) was then added. Four plates were  
131 used in parallel for time-points 15, 30, 60 and 90 min. All plates were placed in the  
132 dark allowing 15, 30, 60 or 90 min, respectively, for dye uptake. Dye and medium  
133 were quickly removed from the plates after incubation and washed with 150 µL

134 fixative solution (1% formaldehyde, 1% calcium chloride) for 2 min. Plates were then  
135 rapidly drained, followed by addition of 200µl extraction buffer (1% acetic acid and  
136 50% ethanol) and left in the dark for 20 min at room temperature. Absorbance of  
137 extracted dye was measured using a microplate reader (Elx808iu Ultra Microplate  
138 Reader, Bio-Tek Instrument Inc., Potton UK) at a wavelength of 570 nm.

139

140

### 141 3. Results and Discussion

#### 142 3.1. Neutral red retention time assay of metal oxide nanoparticles

143 Haemolymph from *M. edulis* was exposed to a panel of metal or metal oxide  
144 nanoparticles at a final concentration of 2ppm (Fig. 1). Lysosomal membrane stability  
145 was tested by measuring NRRT at four different time points; 15, 30, 60 and 90 min.  
146 Results were analysed and statistically compared to the control group using a one-way  
147 anova test with confidence limit of 95% (Figure 2). Lysosomal membrane stability  
148 showed a significant decrease ( $p < 0.05$ ) upon exposure to copper, cobalt and  
149 chromium nanoparticles at all time-points tested, indicating toxic effects on  
150 lysosomes of these nanomaterials. However, no significant effects were observed on  
151 exposure of titanium or gold nanoparticles, suggesting they are less toxic by the  
152 criterion of this *in vitro* assay.

153

#### 154 3.2. Toxicity of metal or metal oxide nanoparticles

155 The particles selected for this study have previously been reported to display a  
156 range of toxicity in biological systems. Titanium dioxide nanoparticles (which are  
157 widely used commercially as a component of sunscreens) are generally regarded as  
158 less toxic to aquatic species (Federici et al, 2007). However, it should be noted that, in  
159 mice, NO and tumour necrosis factor alpha production were elicited after exposure to  
160 titanium dioxide nanoparticles (<10nm). This finding suggested that both damage to

161 the cell structure and macrophage dysfunction may occur, leading to reduction in both  
162 non-specific and specific immune responses in some individual animals (Liu et al  
163 2010). Copper oxide and chromium oxide nanoparticles are notorious for their toxic  
164 effects, and have been implicated in toxicity to non-target organisms (Ivask et al,  
165 2014), reduction of immune status (Zha et al 2009), damage to animal tissues (Chen et  
166 al, 2006; Griffitt et al, 2007), and induction of reactive oxygen species (Fahmy and  
167 Cormier, 2009; Horie et al 2011). Cobalt oxide nanoparticles readily enter cultured  
168 human cells where they are found to have a negative effect on cell viability (Papis et  
169 al., 2009). They have been reported to induce primary DNA damage in a  
170 concentration-dependent manner. Various redox enzyme activities were decreased  
171 after treatment with cobalt nanoparticles, suggesting potential toxic risk and inhibition  
172 of antioxidant capacity (Jiang et al, 2012).

173

### 174 *3.3.Potential for high-throughput assay*

175 The assay format reported here includes minimisation of biological variation in  
176 haemocyte populations by pooling haemolymph across five individual animals.  
177 Moreover, three independent replicates gave essentially identical results and allowed  
178 reproducible discrimination across the nanoparticle panel studied. Use of 96-well  
179 microtitre plates makes possible high-throughput analysis of large numbers of  
180 samples, replicates and concentrations within the time-scale suggested by Moore et al.

181 (2009). This could facilitate rapid quantitative analysis of novel engineered  
182 nanoparticles. An especially attractive feature of this assay format is that it mimics the  
183 kinds of strategies that many nanoparticles most probably employ in nature to gain  
184 entry to cells such as phagocytosis or macro-endocytosis. This is an ancient and long-  
185 established property of eukaryote cells (Elsaesser and Howard, 2012).

186

### 187 **Acknowledgement**

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190

191 **References**

- 192 Babich, H., Borenfreund, E. 1990. Applications of the neutral red cytotoxicity assay  
193 to invitro toxicology. *ATLA - Alt. Lab. Anim.* 18, 129-144.
- 194 Baun, A., Hartmann, N.B., Grieger, K., Kusk, K.O. 2008. Ecotoxicity of engineered  
195 nanoparticles to aquatic invertebrates: a brief review and recommendations for future  
196 toxicity testing. *Ecotoxicol.* 17, 387-395.
- 197 Borenfreund, E., Puerner, J.A. 1985. Toxicity determined *in vitro* by morphological  
198 alterations and neutral red absorption. *Toxicol. Lett.* 24, 119-124.
- 199 Buzea, C., Pacheco, I.I., Robbie, K. 2007. Nanomaterials and nanoparticles: sources  
200 and toxicity. *Biointerphases.* 2, 17-71.
- 201 Canesi, L., Ciacci C., Fabbri, R., Marcornini, A., Pojano, G., Gallo, G. 2012. Bivalve  
202 molluscs as a unique target group for nanoparticle toxicity. *Mar. Environ. Res.* 76, 16-  
203 21.
- 204 Chen, Z., Meng, H., Xiang, G., Chen, C., Zhao, Y., Jia, G., Wang, T., Yuan, H., Ye,  
205 C., Zhao, F., Chai, Z., Zhu, C., Fang, X., Ma, B., Wan, L. 2006. Acute toxicological  
206 effects of copper nanoparticles *in vivo*. *Toxicol. Lett.* 163, 109-120.
- 207 Defosse, J.M., Hawkins, A.J.S. 1997. Selective feeding in shellfish: Size-dependent  
208 rejection of large particles within pseudofeces from *Mytilus edulis*, *Ruditapes*  
209 *philippinarum* and *Tapes decussatus*. *Mar Biol.* 129, 139-147.

- 210 Elsaesser, A., Howard, C.V. 2012. Toxicology of nanoparticles. *Adv. Drug Deliv.*  
211 *Rev.* 64, 129-137.
- 212 Fahmy, B., Cormier, S.A. 2009. Copper oxide nanoparticles induce oxidative stress  
213 and cytotoxicity in airway epithelial cells. *Toxicol. in Vitro.* 23, 1365-1371.
- Federici, G., Shaw, B.J., Handy, R.D. 2007. Toxicity of titanium dioxide nanoparticles to rainbow trout (*Oncorhynchus mykiss*): Gill injury, oxidative stress, and other physiological effects. *Aquat. Toxicol.* 84, 415-430.
- Griffitt, R.J., Weil, R., Hyndman, K.A., Denslow, N.D., Powers, K., Taylor, D., Barber, D.S. 2007. Exposure to copper nanoparticles causes gill injury and acute lethality in zebrafish (*Danio rerio*). *Environ. Sci. Technol.* 41, 8178-8186.
- Gwinn, M.R., Vallyathan, V. 2006. Nanoparticles: health effects--pros and cons. *Environ. Health Perspect.* 114, 1818-1825.
- 214 Hoet, P.H., Bruske-Holfeld, I., Salata, O.V. 2004. Nanoparticles - known and  
215 unknown health risks. *J. Nanobiotechnol.* 2, 12-15.
- 216 Horie, M., Nishio, K., Endoh, S., Kato, H., Fujita, K., Miyauchi, A., Nakamura, A.,  
217 Kinugasa, S., Yamamoto, K., Niki, E., Yoshida, Y., Iwahashi, H. 2011.  
218 Chromium(III) oxide nanoparticles induced remarkable oxidative stress and apoptosis  
219 on culture cells. *Environ. Toxicol.* 28, 61-75.
- 220 Ivask, A., Juganson, K., Bondarenko, O., Mortimer, M., Aruoja, V., Kasemets, K.,  
221 Blinova, I., Henilaan, M., Slaveykova, V., Kahru, A. 2014. Mechanisms of toxic

- 222 action of Ag, ZnO and CuO nanoparticles to selected ecotoxicological test organisms  
223 and mammalian cells *in vitro*: A comparative review. *Nanotoxicol.* 8, 57-71.
- 224 Jiang, C.J., Jia, J.B., Zhai, S.M. 2014. Mechanistic understanding of toxicity from  
225 nanocatalysts. *Int. J. Mol. Sci.* 15, 13967-13992.
- 226 Jiang, H., Liu, F., Yang, H., Li, Y. 2012. Effects of cobalt nanoparticles on human T  
227 cells *in vitro*. *Biol. Trace Elem. Res.* 146, 23-29.
- 228 Klaper, R., Arndt, D., Bozich, J., Dominguez, G. (2014) Molecular interactions of  
229 nanomaterials and organisms: Defining biomarkers for toxicity and high-throughput  
230 screening using traditional and next-generation sequencing approaches. *Analyst* 139,  
231 882-895.
- 232 Liu, R., Zhang, X., Pu, Y., Yin, L., Li, Y., Zhang, X., Liang, G., Li, X., Zhang, J.  
233 2010. Small-sized titanium dioxide nanoparticles mediate immune toxicity in rat  
234 pulmonary alveolar macrophages *in vivo*. *J. Nanosci. Nanotechnol.* 10, 5161-5169.
- 235 Lowe, D.M., Fossato, V.U., Depledge, M.H. 1995. Contaminant induced lysosomal  
236 membrane damage in blood cells of mussels *M. galloprovincialis* from the Venice  
237 Lagoon: An *in vitro* study. *Mar. Ecol. Prog. Ser.* 129, 189-196.
- 238 Lyons, C., Dowling, V., Tedengren, M., Hart, M.G.J., O'Brien, N.M., van Pelt,  
239 F.N.A.M., O'Halloran, J., Sheehan, D. 2003. Immunoblotting determination of levels  
240 of heat shock protein and glutathione S-transferase in Blue mussel, *Mytilus edulis*,



- 241 sampled from Cork Harbour, Ireland, the North and Baltic Seas. Mar. Environ. Res.  
242 56, 585-597.
- 243 Moore, M.N. 1985. Cellular responses to pollutants. Mar. Pollut. Bull.16, 134-139.
- 244 Moore, M.N. 2006. Do nanoparticles present ecotoxicological risks for the health of  
245 the aquatic environment? Environ. Internat. 32, 967-976.
- 246 Moore, M.N., Viarengo, A., Donkin, P., Hawkins, A.J.S. 2007. Autophagic and  
247 lysosomal reactions to stress in the hepato- pancreas of blue mussels. Aquat. Toxicol.  
248 84, 80-91.
- 249 Moore, M.N., Readman, J.A.J., Readman, J.W., Lowe, D.M., Frickers, P.E., Beesley,  
250 A. 2009. Lysosomal cytotoxicity of carbon nanoparticles in cells of the molluscan  
251 immune system: An *in vivo* study. Nanotoxicol. 3, 40-45.
- 252 Nel, A.E., Madler, L., Velego, D., Xia, T., Hoek, E.M.V., Somosundaran, P.,  
253 Klaessig, F., Castranova, V., Thomson, M. 2009. Understanding biophysicochemical  
254 interactions at the nano-bio interface. Nat. Mat. 8, 543-557.
- 255 Nel A, Xia T, Meng H, Wang X, Lin SJ, Ji ZX, Zhang HY. 2013. Nanomaterial  
256 testing in the 21<sup>st</sup> Century: Use of a predictive toxicological approach and high-  
257 throughput screening. Account. Chem. Res. 46, 607-621.

- 258 Papis, E., Rossi, F., Raspanti, M., Dalle-Donne, I., Colombo, G., Milzani, A.,  
259 Bernardinin, G., Gornati, R. 2009. Engineered cobalt oxide nanoparticles readily enter  
260 cells. *Toxicol. Lett.* 189, 253-259.
- 261 Piola, L., Fuchs, J., Oneto, M.L., Basack, S., Kesten, E., Casabe, N. (2013)  
262 Comparative toxicity of two glyphosate-based formulations to *Eisenia Andrei* under  
263 laboratory conditions. *Chemosphere* 91, 545-551.
- 264 Regoli, F. 1992. Lysosomal responses as a sensitive stress index in biomonitoring  
265 heavy-metal pollution. *Mar. Ecol. Prog. Ser.* 84, 63-69.
- 266 Rickwood, C.J., Galloway, T.S. 2004. Acetylcholinesterase inhibition as a biomarker  
267 of adverse effect: a study of *Mytilus edulis* exposed to the priority pollutant  
268 chlorfenvinphos. *Aquat. Toxicol.* 67, 45-56.
- 269 Svendsen, C., Spurgeon, D.J., Hankard, P.K., Weeks, J.M. 2004. A review of  
270 lysosomal membrane stability measured by neutral red retention: Is it a workable  
271 earthworm biomarker. *Ecotox. Environ. Safe.* 57, 20-29.
- 272 Taurozzi, J.S., Hackley, V.A., Wiesner, M. 2010. Preparation of nanoparticle  
273 dispersions from powdered material using ultrasonic disruption. CEINT website. 1-10.  
274 [http://www.nist.gov/customcf/get\\_pdf.cfm?pub\\_id=905633](http://www.nist.gov/customcf/get_pdf.cfm?pub_id=905633).
- 275 Tedesco, S., Doyle, H., Redmond, G., Sheehan, D. 2008. Gold nanoparticles and  
276 oxidative stress in *Mytilus edulis*. *Mar. Environ. Res.* 66, 131-133.

- 277 Tedesco, S., Sheehan, D. 2010. Nanomaterials as Emerging Environmental Threats.  
278 *Curr. Chem. Biol.* 4, 151-160.
- 279 Ward, J.E., Kach, D.J. 2009. Marine aggregates facilitate ingestion of nanoparticles  
280 by suspension-feeding bivalves. *Mar. Environ. Res.* 68, 137-142.
- 281 Widdows, J., Donkin, P. 1992. Mussels and environmental contaminants:  
282 bioaccumulation and physiological aspects. In, *The Mussel Mytilus*. Elsevier Press,  
283 Amsterdam.
- 284 Zha, L., Zeng, J., Sun, S., Deng, H., Luo, H., Li, W. 2009. Chromium(III)  
285 nanoparticles affect hormone and immune responses in heat-stressed rats. *Biol. Trace*  
286 *Elem. Res.* 129, 157-169.
- 287
- 288

289 **Figure legends**

290 **Figure 1** Schematic overview of NRTT assay.

291 **Figure 2** Neutral red retention time (NRRT) assay in response to a panel of  
292 nanoparticles. Neutral red dye extracted from exposed haemocytes was measured  
293 spectrophotometrically at 570nm in a plate reader (\*p< 0.05 versus control values).

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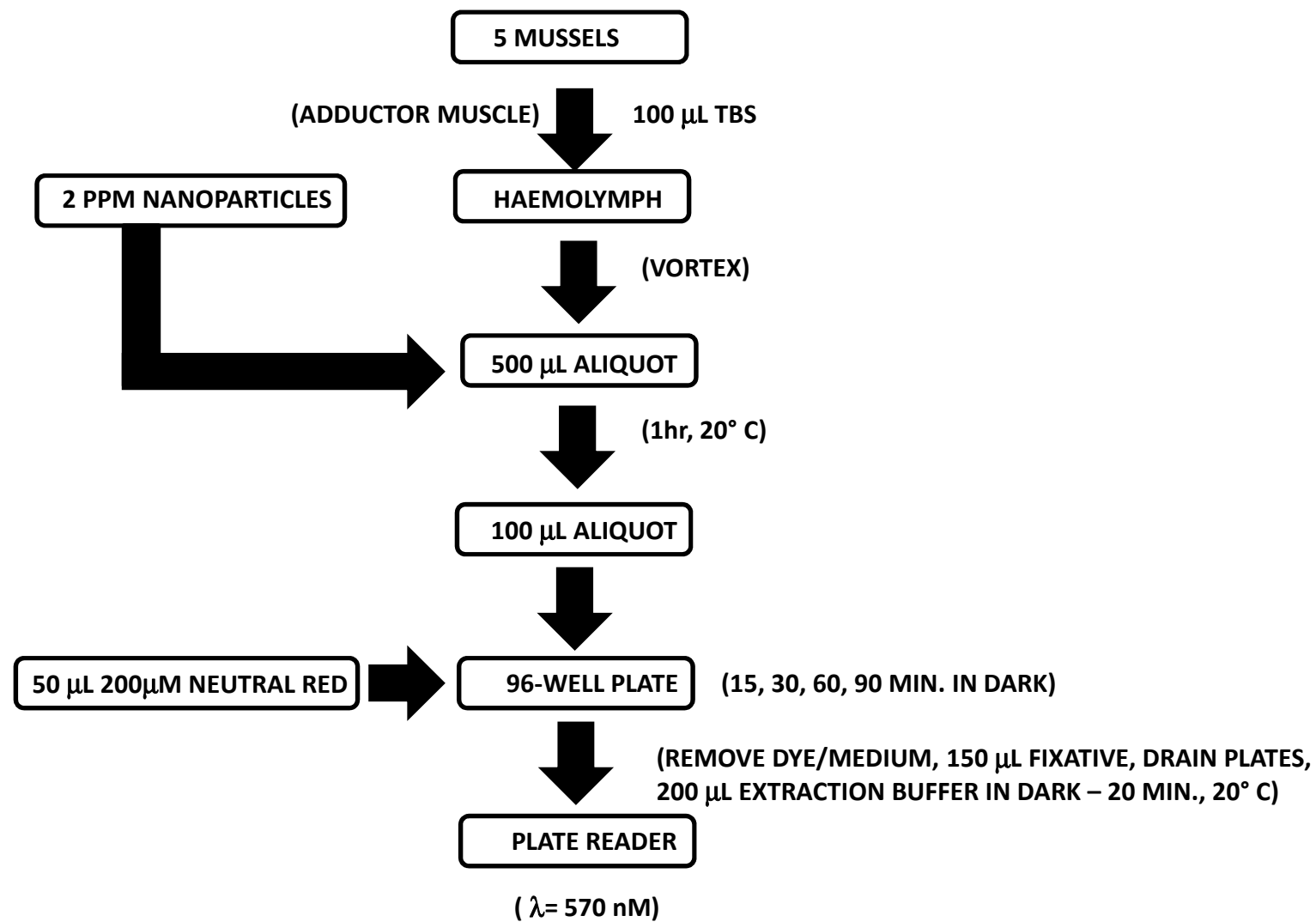


Fig. 1

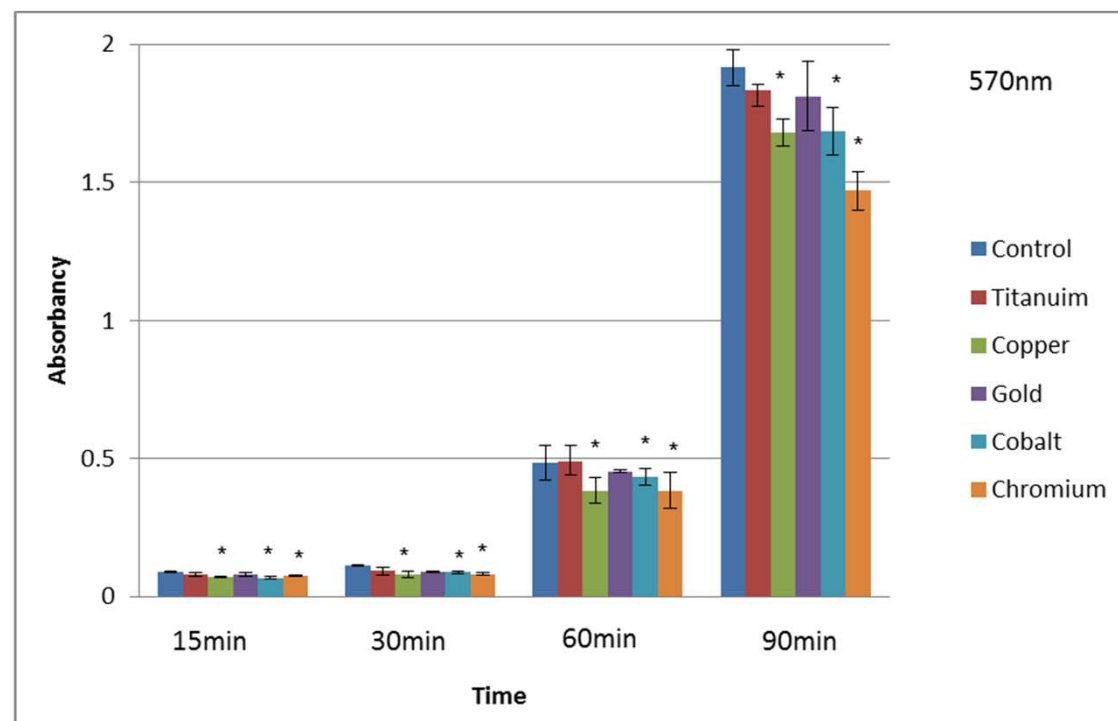


Figure 1

Figure 2

- Neutral red retention time assay used haemolymph of five pooled mussels.
- Assay was miniaturised for reading in a plate reader, facilitating many samples and replicates.
- Copper, chromium and cobalt nanoparticles were toxic while gold and titanium were not.

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