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Geogenic arsenic and uranium in Germany: large-scale distribution control in sediments and groundwater

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Abstract

Arsenic (As) and uranium (U) are naturally occurring trace elements with potentially adverse effects on human health. This work revisits nine case studies on As/U accumulation and remobilization mechanisms in aquifers with different geological and stratigraphical backgrounds to develop a systematic overview of Germany's geogenic inventory of these trace elements. It uses geochemical proxies for a total of 270 solid samples to explain their spatio-temporal distribution: while Pleistocene geological development can explain their extensive absence in sediments and related groundwater in northern Germany, their abundance and distribution in the central and southern parts are widely controlled by sediment provenance geochemistry. Only highly felsic origin (Moldanubian Variscides) enables creation of elevated U in the systems while lower degrees of provenance felsicity (Renohercynian Variscides) appear to be sufficient for As presence. Postdepositional (hydro)geological and anthropogenically triggered intra-basinal processes of trace element accumulation, redistribution and eventually remobilization to groundwater contribute to the present-day situation. Therefore, the ultimate control of these incompatible trace elements is magmatic, even in old sedimentary systems, and still clearly traceable in nowadays large-scale geogenic As and U distribution in Germany and probably elsewhere.

Keywords

magmatic differentiation, geochemical proxies, sediment provenance, Variscides, accumulation processes

1 Introduction

1.1 Arsenic and uranium – a short geochemical overview

Showing characteristics of both chalcophile and siderophile behaviour, arsenic (As) tends to be preferably hosted by sulphide minerals like pyrite or (hydr)oxidic Fe phases like goethite, both of which can contain As up to several wt.% (Smedley and Kinniburgh 2002). In spite of its low average abundance in the upper earth's crust ($1.5\text{--}2\ \mu\text{g g}^{-1}$; Matschullat 1999, Taylor 1964: $1.8\ \mu\text{g g}^{-1}$ in the continental crust), As can accumulate in rocks to concentrations several orders of magnitude higher than these values. The metalloid's fate in the environment is controlled by the prevailing physico-chemical conditions and the presence of other ions. Redox milieu, pH and ionic competition are crucial parameters governing As behaviour (adsorption, desorption, transport, redox transformation). Reducing conditions can lead to As mobilization from oxides while oxidizing conditions may mobilize As bound to sulphides. High groundwater pH constrains As adsorption to mineral surfaces and may therefore be responsible for elevated concentrations in solution. Ions competing with As species for surface binding sites, especially phosphate, can lead to the same result (e.g. Stollenwerk 2002, Smedley and Kinniburgh 2002, Pedersen et al. 2006).

The mobility of U in the environment is, akin to that of As, governed by the Eh-pH milieu and the presence of adsorbers like Fe (hydr)oxides, clay minerals or organic matter (e.g. Doi et al. 1975, Giblin et al. 1981, Merkel and Sperling 1998, Missana et al. 2003, Bots and Behrends 2008). Nevertheless, contrary to As, U is significantly more mobile in its oxidized form U(VI) which is reasoned by its affinity to form stable uranyl hydroxo or calcium uranyl carbonate complexes (e.g. Katsoyiannis et al. 2007, Stewart et al. 2010). In spite of their differing redox and transport properties, As and U frequently occur together in (ground)water of affected areas (e.g. Brown et al. 2007, Kipp et al. 2009, Nicolli et al. 2010, Banning et al. 2012, Banning and Rude 2015).

Taylor (1964) cites the average uranium (U) abundance in the continental crust as $2.7\ \mu\text{g g}^{-1}$. Thereby, generally higher concentrations are detected in felsic rocks (granite average: $4.8\ \mu\text{g g}^{-1}$ U) compared to mafic lithologies (basalt average: $0.6\ \mu\text{g g}^{-1}$ U). Uranium has a strongly incompatible behavior in silicate magmatic differentiation because of its large ion radius and high valence. It is preferentially fractionated into high-temperature metaluminous melts during partial melting and crystal fractionation (Cuney 2010), and therefore often accumulates in granitic and pegmatitic lithologies. Also As is considered one of the incompatible elements which do not fit easily into the lattices of rock-forming minerals precipitating from melt (Webster and Nordstrom 2003).

1.2 Arsenic and uranium as contaminants – health aspects and affected areas

The recognition of As toxicity and its initially underestimated impact on human health on a global scale substantially increased the intensity of As research during the past decades, focussing on large problem areas, especially in Southeast Asia (e.g. Berg et al. 2001, Ravenscroft et al. 2005, Zahid et al.

2009). Consequently, drinking water threshold values were broadly lowered in the 1990ies, mostly down to $10 \mu\text{g L}^{-1}$. This confronted water suppliers with the problem of an increased need to process raw water in order to match drinking water requirements. In recent years, it was also found that in the large majority of cases, naturally occurring As is responsible for elevated groundwater concentrations. Thereby, mobilization from As-enriched minerals is the dominating process (e.g. Lowers et al. 2007). Besides the large tropical, mainly deltaic regions with As-exposed populations (e.g. Bangladesh, West Bengal, Vietnam, Taiwan), (semi)arid As problem areas have been identified worldwide. Prominent examples can be found in Chile (e.g. Oyarzun et al. 2004) and the western U.S.A. (e.g. Welch and Lico 1998). Identified As problem areas in Mexico include the Zimapán Valley (Armienta et al. 2001), the Rioverde Basin (Planer-Friedrich et al. 2001) and the Villa de Reyes Graben around San Luis Potosí (Banning et al. 2012). Even endemic As poisoning was described in the Lagunera region (Del Razo et al. 1990). An overview of elevated As occurrences in Mexico is given by Armienta and Segovia (2008).

While the toxicity of As is well documented and drinking water limitations are established and reconsidered for several decades (actually $10 \mu\text{g L}^{-1}$, WHO 2006), U was neglected in this respect for a long time. It was shown that the risk of U exposure primarily derives from its toxicity as a nephrotoxic heavy metal (i.e., leading to kidney diseases), rather than from its radioactive character (Zamora et al. 1998, Orloff et al. 2004, Bjørklund et al. 2017). For adults in Germany, a total radiation exposure of 2.1 mSv a^{-1} was determined, whereby exposure via drinking water only contributes 0.009 mSv a^{-1} (BfS 2009), i.e. $\sim 0.4 \%$ of the annual radiation dosage. There is no general agreement on fixed limitations for U concentrations in drinking water up to date, although drinking water is considered the most important source of U uptake. WHO announced a “provisional guideline value” of $30 \mu\text{g L}^{-1}$, German legislation decided on a fixed limitation of $10 \mu\text{g L}^{-1}$, valid since 2011. Compared to As, the number of identified geogenic groundwater U problem areas worldwide is low but growing. Nevertheless, it has been recognized that high-U aquifers represent a phenomenon of global extent. Felsic magmatic aquifers in Scandinavia are partly affected by elevated concentrations (Frengstad et al. 2000 and citations therein). Sherman et al. (2007) found 29 % of the studied water samples from a sandstone aquifer in Michigan in excess of the WHO guideline. These authors also give a short overview of case studies from other affected states and countries. In the course of a national monitoring of domestic well water in the U.S.A., Focazio et al. (2006) detected 4 % of all samples exceeding $30 \mu\text{g L}^{-1}$ U. In the north-west Indian state of Punjab, high U concentrations in groundwater have relatively recently been observed over a wide geographical area. Several studies (Patnaik et al. 2015; Kumar et al. 2016; Bajwa et al. 2017; Lapworth et al. 2017) point to potential risks for the exposed population. Unverified hypotheses on the origin of groundwater U include fly ash from coal combustion or input from agricultural fertilisation. However, most authors consider geogenic processes more likely, primarily the weathering of the Siwaliks, i.e. Himalayan foothills (e.g., Patnaik et al. 2015). Also high As groundwaters are known from the region (e.g., Kumar and Singh 2020).

1.3 Known cases of elevated As and U contents in German sediments and groundwater

While the mechanism of large-scale As accumulation in delta sediments of Southeast Asia (e.g. Acharyya et al. 2000, Stanger 2005) or in (semi)arid Latin America (e.g. Bundschuh et al. 2004, Nicolli and Bundschuh 2010) are subject to extensive research and discussions, relatively little was known about the origin and development of geogenic As accumulations in Germany. These have been studied on a rather local to regional scale, although elevated As concentrations in groundwater have been detected in many parts of the country, an overview is given by Heinrichs and Udfluft (1996). Figure 1 illustrates a survey of conducted studies on elevated As occurrences in sedimentary rocks and partly groundwater, whereby no claim to completeness is made. It does neither include high As observed in thermal waters known from e.g., Baden-Baden (Rüde 1996), Aachen (Herch 1997) and Wiesbaden (Rosenberg et al. 1999), nor anthropogenically induced As problems (e.g., from former mining, past agricultural techniques or ammunition). The As-related studies shown in Fig. 1 will be shortly summarized in the following.

Banning et al. (2009) detected partly high As contents up to $140 \mu\text{g g}^{-1}$ in secondary Fe concretions in oxidized Upper Cretaceous (Santonian) sediments from the Münsterland Cretaceous Basin (area 2 in Fig. 1) and identified them as the source for significant As contents in soils of the region. A postdepositional paleo redox event during the late Tertiary caused extremely heterogeneous As and other trace element distribution in shallow marine sediments, leading to partly massively enriched secondary concretions. Mainly pyrite-controlled, rather homogeneous As distribution in the original deeper sediment facies changes under formation of highly reactive redox transition zones and distinct paleo redox boundaries into extremely heterogeneous Fe hydroxide-controlled distribution in near-surface sediments (Banning et al. 2013a). The same mechanism applied to Tertiary marine sediments in the Lower Rhine Embayment (area 5; Banning and Rüde 2010): Santonian and Chattian shallow marine sediments exhibit very similar responses to the late Tertiary oxidative redox event, including rock fabric and mineralogical changes, trace element (esp. As) redistribution and remobilization potentials. The developed genetic and geochemical patterns are obviously of general validity for the studied geoenvironment: large-scale redox events alter comparable lithologies and redistribute trace elements hosted therein in the same way, independent of sediment age.

Concentrations of As in groundwater from southern Lower Saxony (area 3) above the drinking water standard were attributed to output from Lower Triassic (“Buntsandstein”) clastic sediments (up to $693 \mu\text{g g}^{-1}$ As) by Goldberg et al. (1995). Mertens (2000) found an average of $108 \mu\text{g g}^{-1}$ As in Upper Cretaceous (Cenomanian) glauconitic sands in the Ruhr Area (area 4) whereby output from the rocks to groundwater was not observed. In a Pliocene aquifer from the Lower Rhine Embayment (area 5), Cremer et al. (2003) detected up to $130 \mu\text{g L}^{-1}$ As in groundwater, attributable to mobilization via pyrite oxidation, probably triggered by anthropogenic NO_3^- input.

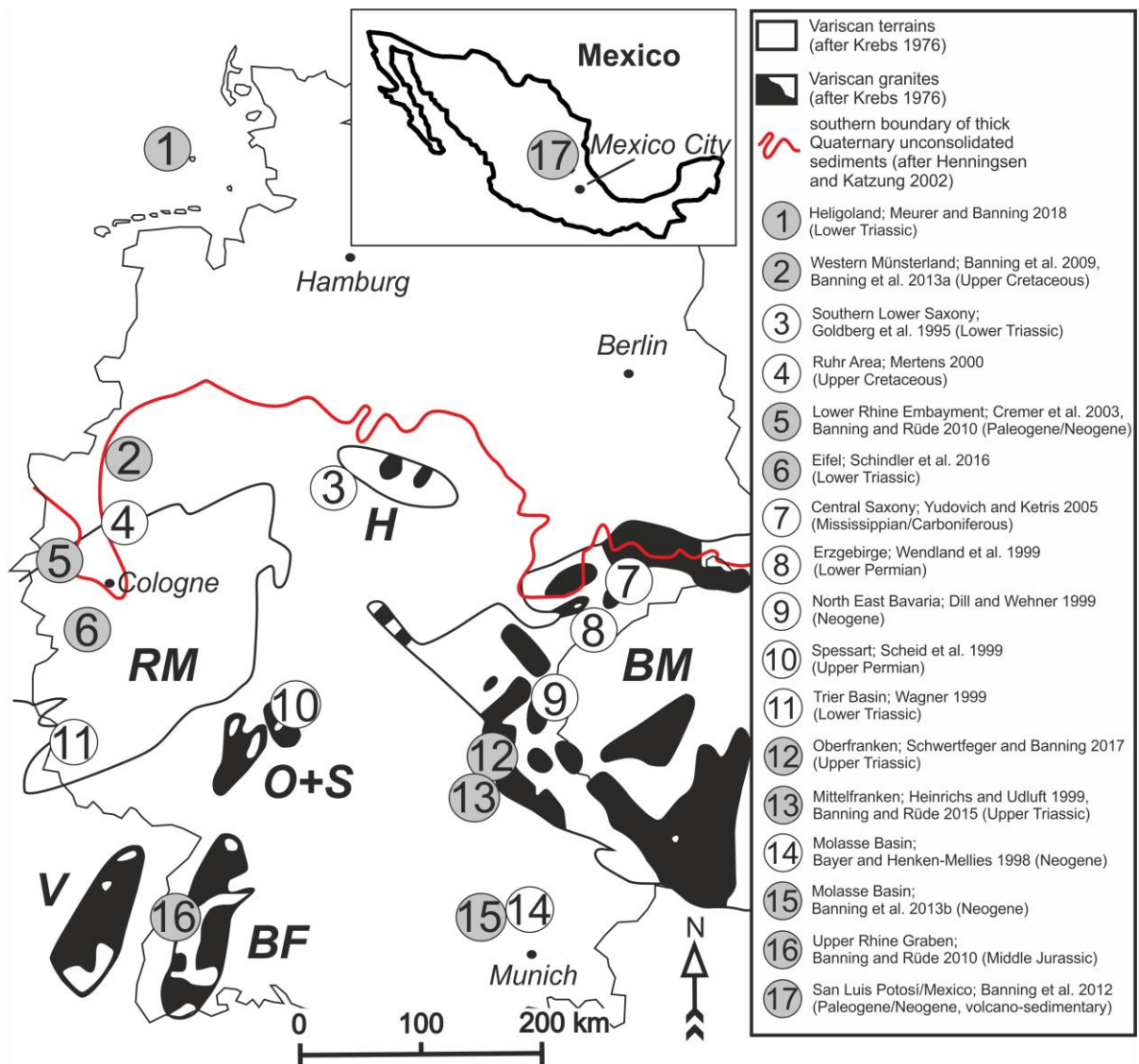


Fig. 1: Distribution of case studies on geogenic As and U in sedimentary rocks and groundwater in Germany (and one study in Mexico; numbering ascending from north to south; studies considered in this work marked as filled grey circles), Variscan terrains and granite intrusions therein (after Krebs 1976) and southern boundary of thick Quaternary cover sediments (after Henningsen and Katzung 2002). Variscan terrains from north to south: H: Harz Mountains, RM: Rhenish Massif, BM: Bohemian Massif, O+S: Odenwald and Spessart, BF: Black Forest, V: Vosges.

Arsenic contents are partly elevated in coal, an overview is given by Yudovich and Ketris (2005). Therein, the authors mention contents of up to $400 \mu\text{g g}^{-1}$ As in Mississippian (Lower Carboniferous) bituminous coals from Saxony (area 7). Interestingly, they found that Eastern German black coal is significantly more enriched than the Western German Upper Carboniferous coal with average As contents of $6.8 \mu\text{g g}^{-1}$. For comparison: the Clarke value for bituminous coal is $9.0 \pm 0.8 \mu\text{g g}^{-1}$. The As host in coal is dominantly pyrite but also organic As can be of importance (Yudovich and Ketris 2005). Wendland et al. (1999) determined a mean As concentration of $55 \mu\text{g g}^{-1}$ in Lower Permian ("Rotliegend") basin sediments in Saxony (area 8) and correlated elevated values to groundwater As anomalies. In Miocene lignite from northeastern Bavaria (area 9), Dill and Wehner (1999) detected

89 $\mu\text{g g}^{-1}$ As on average (Clarke value for lignite: $7.4 \pm 1.4 \mu\text{g g}^{-1}$, Yudovich and Ketris 2005). In single wells in the Spessart (northwestern Bavaria, area 10), a correlation between elevated groundwater As and the distribution of Upper Permian (“Zechstein”) sediments was observed by Scheid et al. (1999). Lower Triassic (“Buntsandstein”) sandstones containing plant fossils in the Trier Basin (area 11) were shown to partly have high As concentrations of up to $880 \mu\text{g g}^{-1}$ (Wagner 1999). Heinrichs and Udluft (1999) found that the distribution of extensively elevated As concentrations in Middle Franconian (northern Bavaria, area 13) groundwater (up to $150 \mu\text{g L}^{-1}$) is dependent on depositional aquifer facies: only terrestrial sediments of the Upper Triassic (“Keuper”) seem to produce groundwater As in excess of the drinking water standard. In southern Bavaria, Middle Miocene terrestrial sands (“Obere Süßwassermolasse”, area 14) contain Fe hydroxide concretions with up to $1900 \mu\text{g g}^{-1}$ As that can locally lead to elevated groundwater concentrations when reducing conditions occur (Bayer and Henken-Mellies 1998). Accumulation processes of As in Middle Jurassic (“Dogger”) sedimentary Fe ores in the Upper Rhine Graben/Baden-Wuerttemberg (area 16) were described by Banning and Rüde (2010): shallow marine environments fostering ooidic Fe ore formation provide conditions for syngenetic As accumulation. The studied depositional conditions proved suitable for As enrichment in mainly goethitic Fe ooids during condensed sedimentation. Thereby, As accumulation is preferred over other trace elements.

Studies on sedimentary aquifer U and associated concentrations in groundwater derived from water-rock-interaction in Germany were scarce until about ten years ago. In contrary, the environmental impact of former extensive U mining, especially in Eastern Germany, has been characterized in detail (e.g. Wolkersdorfer 1996, Winkelmann et al. 2001, Baborowski and Bozau 2006). More recently, the inventory of geogenic background U in groundwater attracted more attention due to the new drinking water limitation. Birke et al. (2010) found a median of $0.17 \mu\text{g L}^{-1}$ U in 908 German bottled water samples (maximum: $16 \mu\text{g L}^{-1}$). They statistically found that Triassic sandstone and crystalline basement aquifers (mainly Black Forest) represent the main hosts for elevated groundwater U. Hessian environmental authorities detected concentrations above $10 \mu\text{g L}^{-1}$ U in 2.7% of 965 analyzed wells with a maximum concentration of $86 \mu\text{g L}^{-1}$, elevated values were ascribed to geogenic input from Triassic rocks or Holocene peat deposits (HLUG 2008). For the SW German federal state of Baden-Württemberg, Liesch et al. (2015) found groundwater U concentrations controlled by geology with highest mean values associated to Upper Triassic aquifers. Uraniferous Upper Triassic sediments are also known from northern Bavaria and referred to as “active arkoses” (Abele et al. 1962). These syndiagenetic carbonate fluorapatite inclusions in the Keuper aquifer sandstones contain up to $1070 \mu\text{g g}^{-1}$ U and were found to show structurally (CO_3^{2-} substitution in the crystal structure) and radiatively (α -recoil damage from U decay) enhanced mineral solubility. Extraction experiments indicated U release to groundwater during weathering: apatite alteration was identified as the responsible mechanism for widespread groundwater U concentrations in the region (16% of wells $>10 \mu\text{g L}^{-1}$) (area 13, Banning and Rüde 2015). Further north, the Keuper aquifer system hosts dolomitic inclusions instead of phos-

phates. Mineralogical and extraction data revealed that also this facies has a significant potential to release U to groundwater (area 12, Steffanowski and Banning 2017); the two uraniferous facies control the geogenic U groundwater problem in northern Bavaria. The subsurface of the “Mechernich Triassic triangle” in the far west of Germany consists of Buntsandstein sediments. Groundwater in this Lower Triassic aquifer contains up to $56 \mu\text{g L}^{-1}$ U, with 7% of all samples in the study dataset exceeding the German guideline value (area 6, Schindler et al. 2016). Further studies describe anthropogenic mobilisation of geogenic subsurface U pools by agricultural activity in Mecklenburg-West Pomerania/NE Germany (van Berk and Fu 2016) and southern Bavaria (area 15, Banning et al. 2013b), drinking water treatment in North Rhine-Westphalia (Banning et al. 2017) or managed aquifer recharge on the island of Heligoland (area 1, Meurer and Banning 2018). The latter is basically a Buntsandstein rock in the North Sea, uplifted by salt tectonics. In the otherwise widely homogeneous red sandstones, Cu- and U-bearing cavity fillings and “fish eyes” – round reduction spots, partly with a black U-rich centre – can be found. They represent the source of geogenic U mainly bound in the carbonate fraction. On Heligoland, brackish water with partly elevated U concentrations is treated via reverse osmosis for use as drinking water. Rain water is irrigated to recharge the aquifer, an artificial accumulation of Buntsandstein debris. The fresh water dissolves part of the geogenic U pool in the sandstones, leading to U groundwater peaks during low tide.

This work revisits and extends nine earlier regional studies which unravelled geogenic As and/or U accumulation processes in aquifers with different geological and stratigraphical background, and their timings in geological history. These aquifers were hydrogeochemically, mineralogically and genetically characterized, they are actually or potentially affected by naturally elevated groundwater concentrations of As and/or U. Besides the major importance of the obtained results on a regional scale and global transferability based on comparable conditions between single areas, more may be learned from geochemical comparison of all study areas and their distribution in space and time. Therefore, the impact of sediment provenance and the spatio-temporal development of Europe on geogenic As and U distribution are evaluated in this study. Geochemical proxies derived from so far unpublished trace element data will be assessed to characterize the studied sediments’ provenance, combine the results with the earlier As/U-focused studies, and derive a geodynamic explanation for large-scale natural As and U distribution in Germany.

2 Materials and Methods

A total of 270 sediment samples from nine study areas (eight studies in Germany, one in Mexico, Fig. 1) were considered in the present study, an overview is given in Table 1. Samples were taken from outcrops and boreholes, details about sediment sampling, geological and hydrogeological conditions, as well as regional As and/or U occurrence can be found in the papers cited in Figs. 1 and 2.

Table 1: overview of investigation areas and samples used in this study (in stratigraphical order, cf. Fig. 1 for location of investigation areas).

area no.	n=	stratigraphy	main lithology; As/U host phases
15	46	Neogene/Quaternary	Miocene alluvial sands, Quaternary fluvial gravels; Lignite, Fe hydroxides
5	36	Paleogene/Neogene	little consolidated marine sands, lower reduced and upper oxidized facies; pyrite, Fe hydroxide concretions
17	30	Paleogene/Neogene	terrestrial volcano-sedimentary basin filling (little consolidated sands, playa lake sediments); carbonates, Fe hydroxides
2	38	Upper Cretaceous	little consolidated marine sands, lower reduced and upper oxidized facies; pyrite, Fe hydroxide concretions, siderite concretions
16	32	Middle Jurassic	shallow marine limestone, mudstone, sandstone, ooidic Fe ores; Fe (hydr)oxides
12	14	Upper Triassic	terrestrial sandstone; uraniferous dolomite inclusions
13	47	Upper Triassic	terrestrial sandstone; uraniferous apatite inclusions, Fe oxides
1	18	Lower Triassic	red terrestrial sandstone; carbonate cavity fillings, U-rich black reduction spots
6	9	Lower Triassic	red terrestrial sandstone; Fe hydroxides

All samples were ground to powder grain size and analysed for bulk rock trace element geochemistry using Instrumental Neutron Activation Analysis (INAA). A sample aliquot of 1 g was encapsulated in a polyethylene vial and irradiated along with flux wires at a thermal neutron flux of $7 \cdot 10^{12} \text{ n cm}^{-2} \text{ s}^{-1}$. ^{24}Na was allowed to decay for 7 days. Subsequently, the samples were counted on a high purity Ge detector with resolution of better than 1.7 KeV for the 1332 keV ^{60}Co photopeak. These analyses were performed by ISO 17025:2017-accredited Activation Laboratories Ltd., Ancaster, Ontario/Canada. For the trace element data presented here, analytical detection limits were $0.5 \mu\text{g g}^{-1}$ for La, U and As; $0.2 \mu\text{g g}^{-1}$ for Th and $0.1 \mu\text{g g}^{-1}$ for Sc. For quality control, flux wires and control standards were used to compare decay-corrected activities to a calibration developed from certified international reference materials: DMMAS 108, 108-B, 111 or 119. One standard was run for every 11 samples. Replicates and blanks were analysed to check accuracy and precision of the data.

3 Results and Discussion

3.1 Regional significance and comparative overview

Important outcomes from five of the nine regional studies evaluated in the course of the present work are visualized in Fig. 2 whereby As and U sources and sinks structures including mobilization and immobilization processes in the studied areas are assessed. Studies 13, 16, 2 and 5 mentioned in Fig. 2 were conducted in different parts of Germany (Fig. 1, chapter 1.3). Study 17 (Banning et al. 2012) discovered elevated groundwater As and U concentrations in the Villa de Reyes Graben, San Luis Potosí/central Mexico and succeeded in unravelling their common occurrence as well as the evolution of both elements' geochemical signatures. Absolute concentrations, inter-elemental behaviour and geochemical proxies supported the aforementioned view that both U and As are incompatible elements in magmatic differentiation and therefore accumulated in felsic lithologies. Dissolution of volcanic glass was identified as the main common release mechanism of U and As in the volcano-sedimentary Mexican basin. It represents input from a common source into groundwater, supported by geochemical signatures (e.g., normalized REE patterns) in volcanic rocks, basin fill sediments and groundwater (Banning et al. 2012). While As and U can both occur in groundwater of sedimentary basins, it became obvious that in most samples, only one of the trace elements showed high concentrations. This is explainable by separation due to redox heterogeneity in the aquifer, or by additional preferential input from secondary sedimentary sources (in this case preferential As over U input due to desorption from Fe (hydr)oxides and decarbonatization from caliche deposits).

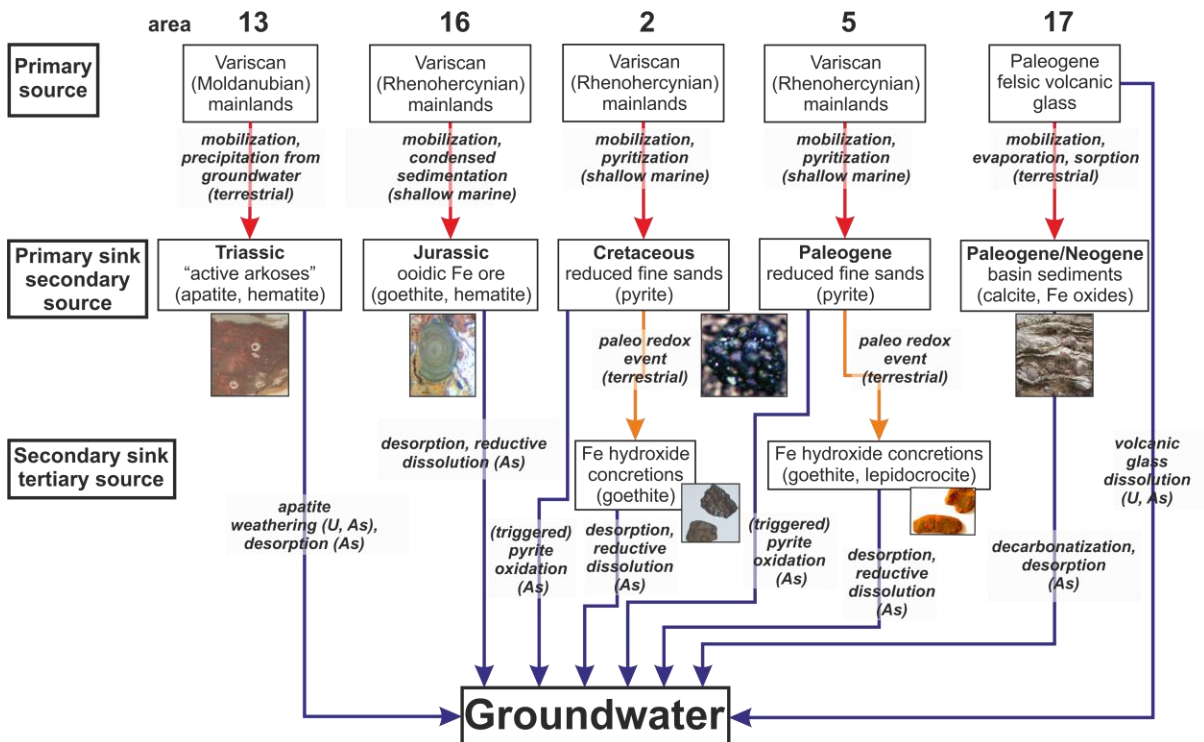


Fig. 2: Graphical overview of As and U sources and sinks structures including mobilization and immobilization processes in some of the studied areas. Area 13: Banning and Rude (2015), area 16: Banning and Rude (2010), area 2: Banning et al. (2013), area 5: Banning and Rude (2010), area 17: Banning et al. (2012).

Figure 2 indicates that trace element accumulation processes and reservoirs can be of different orders such that spatio-temporal structure models of sources and sinks, and transfer processes between them, can be developed for a given area. The choice of study areas representing a variety of geological backgrounds and evolution make approaches and obtained results transferable to other regions. Sandstone-hosted uraniferous phosphates (area 13), shallow marine ooidic Fe ores (area 16), reduced glauconitic/pyritic marine sandy sediments (areas 2 and 5) and felsic volcano-sedimentary basins (area 17) are widespread geological environments worldwide and often share genetic and geochemical characteristics. This is especially supported by striking similarities between the studies in Cretaceous and Paleogene sediments (areas 2 and 5) exhibiting close analogies not only in geological situation and development but also in As distribution and behaviour in spite of a time span of ~50 Ma between sediment deposition in both areas.

3.2 Geodynamic model for large-scale natural As and U distribution in Germany

Besides the major importance of the obtained results on a regional scale and global transferability based on comparable conditions between single areas, more may be learned from general geochemical comparison of all study areas and their distribution in space and time. Therefore, the impact of sediment provenance and the spatio-temporal development of Europe on geogenic As and U distribution are evaluated in the following. Geochemical proxies are assessed to characterize the studied sediments' provenance and combine the results with the As and U accumulation processes derived in earlier studies. Figure 3 illustrates a ternary plot of La-Th-Sc to deduce source rock geochemistry and tectonic setting of the studied sediments.

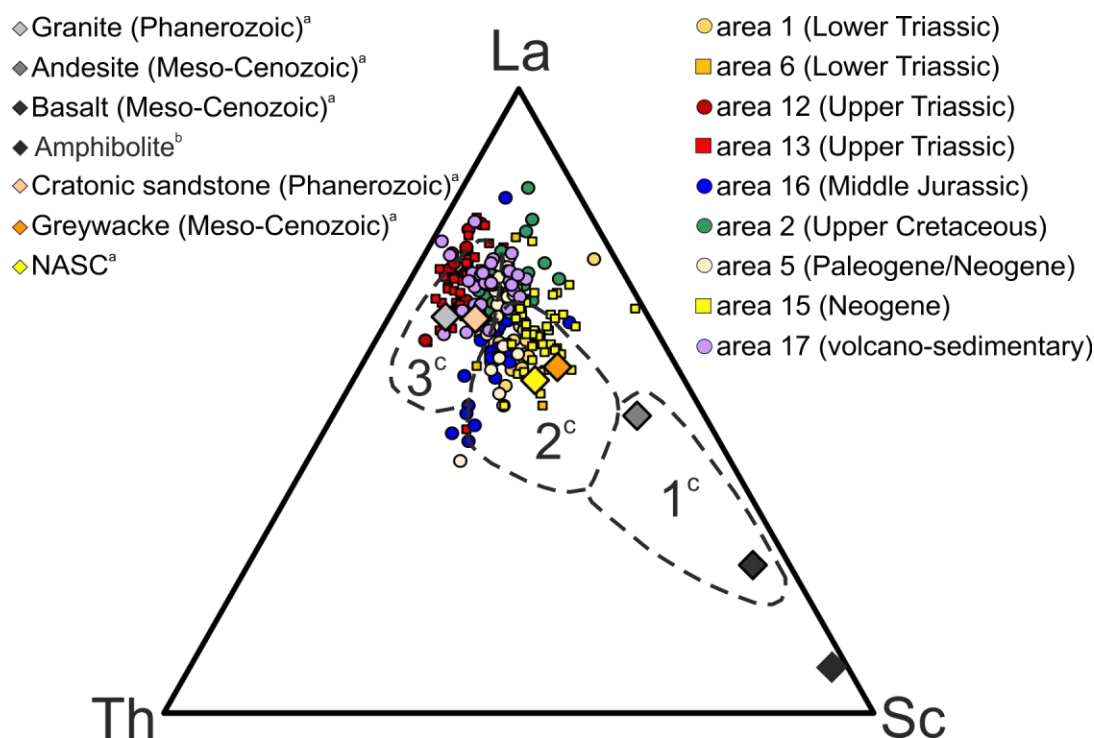


Fig. 3: La-Th-Sc plot of all studied sediment samples (cf. Fig. 1 for study area allocation). For comparison, plots of typical lithologies were implemented after ^aCondie (1993) and ^bCullers (1994). Furthermore, fields in the diagram after ^cBhatia and Crook (1986) indicate provenance rock tectonic setting with 1: oceanic island arc, 2: continental island arc, 3: active or passive continental margin. NASC: North American Shale Composite.

Most studied sediments plot in a rather dense cloud near the La corner of the diagram. Regarding the distal position of the samples from standard amphibolite, basalt and andesite (characterized by higher contents of compatible Sc) and the proximity to granite (higher degree of incompatible La and Th), a generally rather felsic provenance is obvious for all sediments (Fig. 3). Also in comparison to NASC, the North American Shale Composite (representative of upper crustal sediments; e.g., Gromet et al. 1984), the majority of samples plot in the more felsic range and are in this respect rather comparable to typical cratonic sandstone plotting within the data aggregation.

Concerning tectonic setting of the provenance sensu Bhatia and Crook (1986), most samples indicate derivation from continental margins with an overlap to the field of continental island arcs (Fig. 3), thus reflecting the tectonic configuration of Variscan central Europe as will be discussed later on. Sediment origin from oceanic island arcs can be excluded which was expectable regarding the choice of study areas.

To better differentiate between the single study areas with regard to geochemical provenance, a binary plot of incompatible La vs. compatible Sc is presented in Fig. 4.

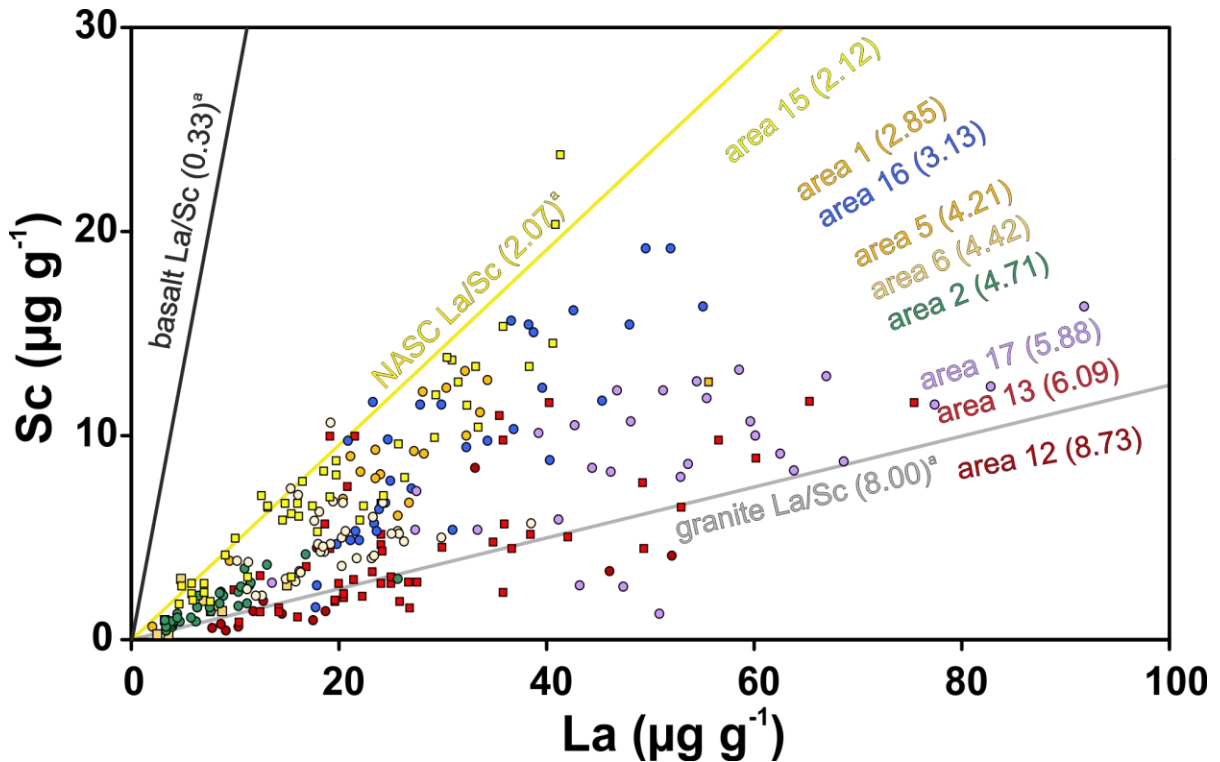


Fig. 4: La vs. Sc scatter plot and average La/Sc ratios for sediment samples from the nine study areas, cf. Fig. 1 for study area allocation. Basalt, NASC and granite standard ratios after ^aCondie (1993).

The observation of geochemical proxies indicating rather felsic provenance, also in comparison to average upper crustal sediment (NASC), is supported by the binary plot (Fig. 4). In spite of partly considerable scattering (R^2 of the nine data subsets: 0.11-0.72), differences between the single study areas become evident. Upper Triassic sediments from areas 12 (La/Sc ratio derived from trendline equation: 8.73) and 13 (6.09) as well as volcano-sedimentary basin filling from San Luis Potosí (area 17, 5.88) exhibit clearly more felsic provenance than marine sediments of Jurassic (area 16), Paleogene (area 5), Cretaceous (area 2) and Lower Triassic (areas 1 and 6) age. Neogene sediments from the southern German Molasse basin (area 15) document average La/Sc ratios very closely to the NASC signature.

Consequently, areas with high U potential in terms of accumulation in rocks/sediments and elevated concentrations in groundwater originate from more felsic sources than sediments hosting “only” As.

The following is hypothesized:

- a) Sedimentary environments potentially fostering As enrichment require above-average (i.e., above NASC) felsic provenance to assure sufficient supply of the incompatible trace element.
- b) Sedimentary environments potentially fostering U enrichment require highly felsic provenance to assure sufficient supply of the (more) incompatible trace element, i.e. highly felsic origin is a prerequisite for sedimentary systems to create both, high As and U. Of course, con-

ditions of redeposition, climate, hydrochemistry and microbiology (let alone anthropogenic activity) control potential final groundwater concentrations.

- c) Classical geochemical proxies like those presented here may serve as pre-diagnostic tools to characterize geochemical provenance (and thus, primary trace element sources) of sedimentary areas potentially affected by As and/or U.

In an effort to combine this approach and the previously discussed As/U accumulation processes, plots of La vs. U and La vs. As are presented as Fig. 5 and Fig. 6, respectively.

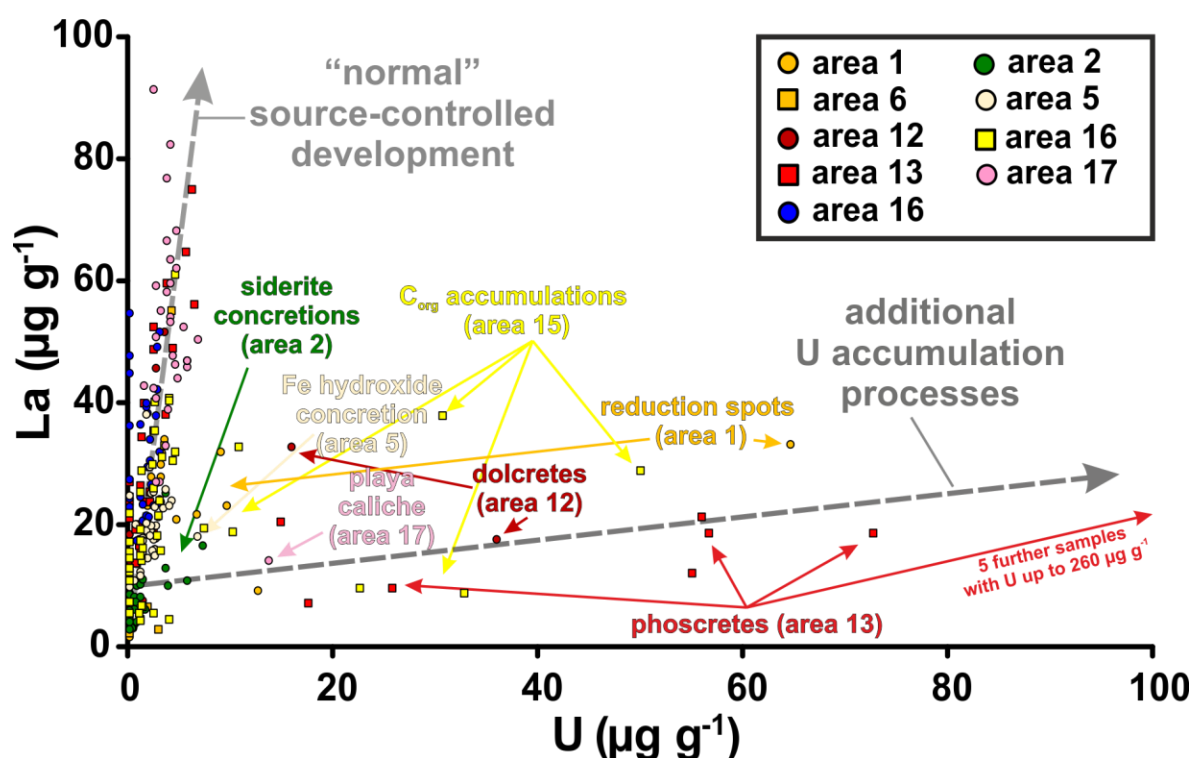


Fig. 5: U vs. La scatter plot for all studied sediments (cf. Fig. 1 for study area allocation), and secondary U accumulation mechanisms in sediment basins (cf. Fig. 2 and chapter 1.3).

The majority of samples in Fig. 5 exhibit a development along a positively trending line interpreted to represent “normal” source (i.e. provenance)-controlled behaviour. Here, common successive increase of both incompatible trace elements is obvious. Offsets from this “background” development, i.e. excess U, reflect additional sedimentary enrichment processes in the different areas. Slight accumulation was found in siderite concretions within Cretaceous fine sands (area 2). Uranium sorption to the purest Fe hydroxide concretion in oxidized sands of the Lower Rhine Embayment (area 5, this sample has the singular highest As content ($1860 \mu\text{g g}^{-1}$) of all study area) was not able to accumulate more than $7.2 \mu\text{g g}^{-1}$ U. This finding supports the hypothesis that a more felsic provenance is needed for sufficient U supply – U availability was just too low in the Paleogene sediments such that not even this concretionary best option for accumulation could be taken advantage of. Evaporative concentration accounts for

elevated U contents in relic playa lake caliches (area 17). Apatite precipitation from U-bearing groundwater on former playa carbonates resulted in high U contents in phoscretes (area 13), also the more proximal original dolcretes (area 12) show some U accumulation. Very localized reducing conditions lead to punctual U hot spots in Lower Triassic sandstone on Heligoland (area 1), this phenomenon is also known from e.g., Switzerland (Hofmann 1990, Burkhalter 1995). Uranium accumulation in Neogene sediments (area 15) is almost exclusively observed in postsedimentary, i.e., Holocene, C_{org} -rich deposits like peat which successively concentrated U from groundwater.

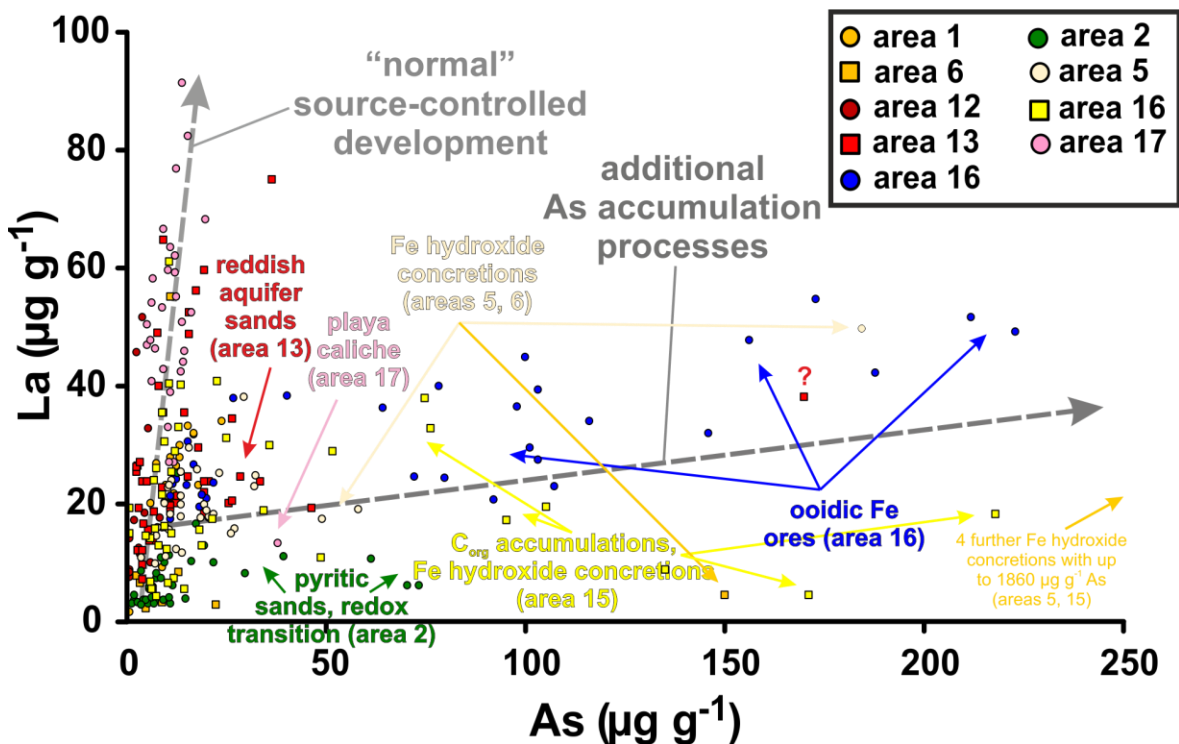


Fig. 6: As vs. La scatter plot for all studied sediments (cf. Fig. 1 for study area allocation), and secondary U accumulation mechanisms in sediment basins (cf. Fig. 2 and chapter 1.3).

Similar observations can be made in the As vs. La plot (Fig. 6) – a provenance-driven “background” development is overlain by different additional As concentration processes occurring in the studied sediment basins: sorption to Fe hydroxide coatings of silicate aquifer material (areas 6 and 13), accumulation in ooidic Fe ores during condensed sedimentation (area 16), fixation in pyrite and subsequent redistribution to redox transition zones and Fe hydroxide concretions (areas 2 and 5), enrichment by evaporation in playa caliches (area 17), concentration by sorption to organic material and in Fe(III) concretions (area 15). Jurassic ooidic Fe ores (area 16) with high As values only yield U contents of $1.5 \mu\text{g g}^{-1}$ on average although depositional environment and potential host phases seem suitable for higher accumulation. Again, U availability in the sediment provenance is suggested to be the limiting factor for this phenomenon. Compared to the U-La plot (Fig. 5), the trend of “normal” source-controlled development between As and La for all samples appears less distinct. Obviously, the signa-

ture of the more incompatible U regarding geochemical provenance proxies is clearer. Moreover, additional redistribution processes after sediment deposition (like paleo redox events) seem to preferentially fractionate As, thereby altering primary geochemical signals.

To pursue these observations and possibly further support the formulated hypotheses, the stratigraphical and spatial distribution of the study areas will be assessed in the following and set in relation to their sediment provenance areas. Figure 7 offers an overview of the study areas' stratigraphical situation and the plate tectonic framework.

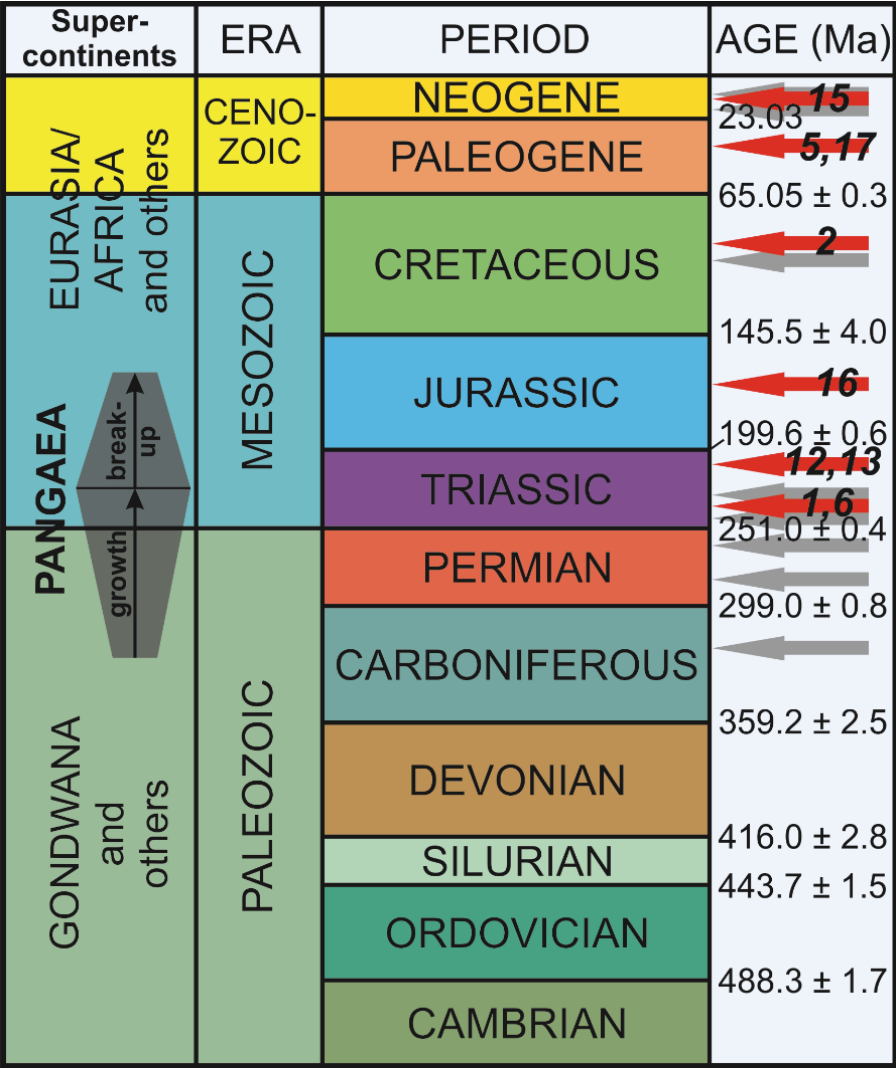


Fig. 7: Stratigraphical distribution of study areas (red arrows) and studied occurrences of sedimentary As in Germany from literature (grey arrows, cf. Fig. 1), absolute stratigraphical boundary ages (right side) after ICS (2009). Additionally, major crustal development (supercontinents, left side) after Bahlburg and Breitzkreuz (1998) is indicated.

The stratigraphical distribution of studied elevated sedimentary As in Germany reflects affected units ranging from Carboniferous to Neogene in age. A clustering of cases in Permo-Triassic is conspicuous which includes Germany's most significant As (and U) problem area in Northern Bavaria (areas 12,

13). Sediments from these periods often represent cratonic sandstones (cf. Fig. 3) derived from weathering of the then young Variscan massifs. Surveys of groundwater U conducted so far identified the Triassic as “hot spot” age for affected aquifers (Birke et al. 2010, Kurth 2010, HLUG 2008, LGL and LfU 2007; cf. chapter 1.3). Likewise, an overview of As in German groundwater (Heinrichs and Udluft 1996) reports elevated concentrations mainly in Permo-Triassic aquifers. During that time, the formation climax of the supercontinent Pangaea as a result of the collision between Laurasia and Gondwana was reached (Bahlburg and Breitzkreuz 1998, Fig. 7). This process was accompanied by the Variscan Orogeny representing a mountain-building event, relics of which are found today as e.g., the Rhenish Massif, the Black Forest or the Bohemian Massif in Germany and neighbouring countries. Denudation of the orogen and transport of resulting material to sediment basins started in the Upper Carboniferous. Concluding from Fig. 7, only stratigraphies younger than that appear to be affected by elevated As. Moreover, the upper limit of a stratigraphical “window” of sedimentary As occurrences seems to be defined by the end of the Neogene period. Possible explanations for that must be sought in the spatio-temporal geological development of Germany.

Figure 1 illustrates the spatial distribution of studied sedimentary As as well as Variscan terrains and thick Quaternary overburden. It becomes evident that all known cases of elevated sedimentary As in Germany are located south of the thick unconsolidated Quaternary cover governing geology and landscape of northern Germany (except for the Triassic rock island Heligoland, area 1, which is not covered by Quaternary). These mostly Pleistocene glacial and periglacial deposits with main provenance areas in northern and northeastern Europe (Henningsen and Katzung 2002) obviously did not bring a significant As (or U) potential along. Not only would these observations explain the northern boundary of As study areas (Fig. 1) but also raise expectations concerning better shallow groundwater quality with respect to As and U in northern Germany. Indeed, Kunkel et al. (2004) characterized natural groundwater quality in Germany and found in an evaluation of a large As dataset (n=1661) from sand and gravel aquifers of the North German Plain (widely identical with the area north of the red line indicated in Fig. 1) 90th percentile values of 3.3-3.5 $\mu\text{g L}^{-1}$ As with hardly any single values above 10 $\mu\text{g L}^{-1}$ in different aquifer depth intervals of up to 50 m below ground surface. Supporting this in their earlier survey, Heinrichs and Udluft (1996) do not report on geogenic groundwater concentrations in exceedance of 10 $\mu\text{g L}^{-1}$ As from this area.

The organisation foodwatch collected and published drinking water U data from German federal state authorities (foodwatch 2009). In the dataset, they found that concentrations in the northern and central federal states Brandenburg, Berlin, Bremen, Hamburg, Schleswig-Holstein, Lower Saxony, North Rhine-Westphalia and Saarland are quantitatively below 10 $\mu\text{g L}^{-1}$ U (2 out of 427 samples in Mecklenburg-Western Pomerania were above that limit) while southern federal states generally exhibit more abundant violations of the drinking water guideline.

Another observation from Fig. 1 is the close spatial relation of most conducted sedimentary As and U studies to Variscan terrains in Germany supporting a relation between Variscan geology and trace

element distribution as was already suspected from stratigraphy (Fig. 7). Differences between sedimentary occurrence of As and U in the single studied areas in this work can be derived from the distribution of Variscan granites and support the hypotheses on degree of provenance felsicity formulated earlier: Upper Triassic sediments from northern Bavaria (areas 12, 13) derive from weathering of the Bohemian Massif which belongs to the Moldanubian section of the European Variscides (like main parts of Black Forest and Vosges) representing the most metamorphic part with abundant felsic intrusions (Krebs 1976, Bahlburg and Breiterkreuz 1998) while Cretaceous and Paleogene sediments (areas 2, 5) originate from the Rhenohercynian Variscan section with low degree of metamorphism and widely absent granitic intrusions (Fig. 1). At first glance, it seems surprising that the Jurassic sediments (area 16), being located in direct vicinity to the Moldanubian granite-rich Black Forest and Vosges (Fig. 1) do neither show highly felsic provenance (Fig. 4) nor elevated U potential (Fig. 5). Nevertheless, the Moldanubian crystalline terrains did not serve as sediment provenance but were submerged during the Middle Jurassic (Ziegler 1990), sediment (and Fe/As, but no U) input was accomplished from the Rhenish Massif (Sauer and Simon 1975) and thus, from the Rhenohercynian – a provenance of lower felsicity.

To crosscheck this hypothesis, a short review of another basin that actually received its sedimentary filling from weathering of the Black Forest was conducted. Within the range of this Moldanubian terrain, only few rather small sedimentary basins filled with Upper Carboniferous to Permian material fulfill this condition, the biggest of which is the one around Baden-Baden in the northern Black Forest (Henningsen and Katzung 2002). Stefanian (Upper Carboniferous) partly C_{org}-rich arkoses and sandstones in that area, underlain by granite and derived from erosion of the proximal granitic young Variscan mountains, host several sedimentary U anomalies prospected in the past, the most significant of which is the deposit Müllenbach (Kneuper et al. 1977). The mineralization is of the roll-front type, main U carriers are uraninite and coffinite, bulk U₃O₈ contents of several wt.% are known. Arsenic is present in concentrations of up to 3000 µg g⁻¹, mainly as arsenopyrite. Kneuper et al. (1977) also found a positive correlation between As and U in the sediments. The authors genetically suggest common U and As precipitation from groundwater with trace elements primarily mobilized from Black Forest granites. This development is thus very similar to the one described here for Triassic phoscretes (area 13). Also Permian sediments in that basin partly show common enrichment of both U and As (Plinninger and Thuro 1999). Hydrothermal U mineralizations within granites are known from the southern Black Forest, e.g., near Menzenschwand. This site is the type locality of several secondary U-As minerals, one of which – Nielsbohrite [(KUO₂)₃(AsO₄)(OH)₄*H₂O] – was described by Walenta et al. (2009). These studies clearly underline Black Forest granites' high source potential for both U and As. The receiving sediment basins in the area are probably too small (in comparison to e.g., the South German Keuper Basin) to account for significant postdepositional redistribution processes and prominent groundwater As and U problems. Nevertheless, elevated U concentrations in Black Forest mineral waters are known (Birke et al. 2010). Moreover, the thermal springs of Baden-

Baden, recharged by meteoric waters circulating in the regional granites, show high As concentrations of 200 $\mu\text{g L}^{-1}$ on average; U presence is documented by U-bearing amorphous thermal Mn hydroxides and opaline sinters (Rüde 1996). In consequence, this short overview of a second Moldanubian massif fully supports the model of Variscan control over geogenic As and U distribution in Germany developed in the present work.

One exception among the existing studies of sedimentary As in Germany probably without Variscan influence should be mentioned: Bayer (1997) states that partly high As concentrations (up to 1900 μg^{-1}) in Fe-rich sections of the Bavarian molasse basin, esp. Badenean sands (area 14 in Fig. 1), originate, like the sediments themselves, from weathering of the Alpine Mountains; As is believed to primarily derive from mineralized zones in the eastern Alps. Uranium concentrations are low in the Fe concretions, Bayer (1997) detected $<1 \mu\text{g g}^{-1}$ U in all his samples. According to the genetic model developed in the present work, this data argues for a primary trace element source of moderately high felsicity not able to supply sufficient U for enrichment in sediments. This source indeed is likely to be found in the eastern Alps which, in contrast to the central and western parts of the orogen, do not host significant granitic intrusions (Gwinner 1971). Nevertheless, several occasions of elevated U are known in groundwater and sediments from the range of the Miocene molasse basin (e.g. LGL and LfU 2007). An explanation for that can be given by another paleogeographical observation: after deposition of the Badenean sands, sediment provenance changed for the eastern and northern parts of the basin. During the subsequent Sarmatian and Pannonian, clastic material and groundwater supply were partly accomplished from the Bohemian Massif (Unger 1989) – a provenance of high felsicity with proven ability to produce U enrichment in associated sediment basins.

The conclusions drawn above may partly be transferable to other countries and even orogenies. In a survey of U occurrences in British groundwater, Smedley et al. (2006) found highest concentrations in sandstones derived from weathering of the Caledonian orogen (e.g. Old Red Sandstone) and speculate on U sorbed to abundant Fe (hydr)oxides as direct source for elevated groundwater U concentrations (no solid phases were studied). The primary source is likely to be the Caledonides which exhibit, akin to the German Moldanubian Variscides, numerous granite intrusions (e.g. Brown and Locke 1979).

Limitations of this study include the partly unequal distribution of the number of samples from the different study areas, ranging from 9 to 47, associated with heterogeneous levels of significance in statistical considerations. Also, trace element rich sample types (concretions, inclusions etc.) are likely to be overrepresented in the dataset and therefore do not resemble the composition of the whole aquifer/sediment system. Lastly, revisited studies were published within a decade (2010-2019), with partly differing people accomplishing sampling, sample preparation, sample analysis.

4 Conclusions

Geological evolution, expressed by geochemical proxies, can explain trace element distribution on different temporal and spatial scales, and help understand and forecast the occurrence of actual and potential groundwater quality problem areas. Trace element abundance was shown to directly reflect supraregional and intra-basinal geological evolution. The distribution of areas with elevated As and U in Germany is large-scale widely determined by Variscan and Quaternary geology. Geochemical sediment provenance controls elevated As (felsic provenance) and U (highly felsic provenance) supply to sedimentary environments, whereby the different Variscan sections are decisive. Thus, magmatic geochemistry based on incompatibility of U and As is the ultimate control of trace element supply to sedimentary systems where subsequent intra-basinal processes of trace element accumulation, redistribution and eventually remobilization to groundwater take place and create the present-day situation.

The present work contributes to a deeper understanding of the interplay between geological history (magmatic and sedimentary), geochemistry, mineralogy, hydrogeology and hydrochemistry with respect to potentially hazardous geogenic trace elements, focussing on the still little understood situation in Germany. Increased comprehension of occurrence and behaviour of not only anthropogenic but also natural contaminants on a global scale will be necessary in the future in view of growing awareness regarding health impacts along with more stringent drinking water limitations, and continuously more severe population pressure on water resources, especially under climate change conditions.

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References

- Abele, G., Berger, K., Salger, M., 1962. Die Uranvorkommen im Bursandstein Mittelfrankens [The uranium occurrences in the Bursandstein of Middle Franconia]. *Geologica Bavarica* 19, 3-90 (in German).
- Acharyya, S.K., Lahiri, S., Raymahashay, B.C., Bhowmik, A., 2000. Arsenic toxicity of groundwater in parts of the Bengal basin in India and Bangladesh: the role of Quaternary stratigraphy and Holocene sea-level fluctuation. *Environmental Geology* 39, 1127-1137.

557 Armienta, M.A., Villaseñor, G., Rodriguez, R., Ongley, L.K., Mango, H., 2001. The role of arsenic-
558 bearing rocks in groundwater pollution at Zimapán Valley, Mexico. *Environmental Geology*
559 40, 571-581.

560 Armienta, M.A., Segovia, N., 2008. Arsenic and fluoride in the groundwater of Mexico. *Environmen-
561 tal Geochemistry and Health* 30, 345-353.

562 Baborowski, M., Bozau, E., 2006. Impact of former mining activities on the uranium distribution in
563 the River Saale (Germany). *Applied Geochemistry* 21, 1073-1082.

564 Bahlburg, H., Breitzkreuz, C., 1998. *Grundlagen der Geologie [Basics of Geology]*. Enke, Stuttgart,
565 328 pp (in German).

566 Bajwa, B.S.; Kumar, S.; Singh, S.; Sahoo, S.K.; Tripathi, R.M. (2017): Uranium and other heavy toxic
567 elements distribution in the drinking water samples of SW-Punjab, India. – *Journal of Radia-
568 tion Research and Applied Sciences* 10: 13-19.

569 Banning, A., Coldewey, W.G., Göbel, P., 2009. A procedure to identify natural arsenic sources, ap-
570 plied in an affected area in North Rhine-Westphalia, Germany. *Environmental Geology* 57 (4),
571 775-787.

572 Banning, A.; Rüdé, T.R., 2010. Enrichment processes of arsenic in oxidic sedimentary rocks – from
573 geochemical and genetic characterization to potential mobility. – *Water Research* 44, 5512-
574 5531.

575 Banning, A.; Cardona, A.; Rüdé, T.R., 2012. Uranium and arsenic dynamics in volcano-sedimentary
576 basins – an exemplary study in north-central Mexico. – *Applied Geochemistry* 27, 2160-2172.

577 Banning, A.; Rüdé, T.R.; Dölling, B., 2013a. Crossing redox boundaries – Aquifer redox history and
578 effects on iron mineralogy and arsenic availability. – *Journal of Hazardous Materials* 262,
579 905-914.

580 Banning, A.; Demmel, T.; Rüdé, T.R.; Wrobel, M., 2013b. Groundwater uranium origin and fate con-
581 trol in a river valley aquifer. – *Environmental Science & Technology* 47, 13941-13948.

582 Banning, A.; Rüdé, T.R., 2015. Apatite weathering as a geological driver of high uranium concentra-
583 tions in groundwater. – *Applied Geochemistry* 59, 139-146.

584 Banning, A.; Pawletko, N.; Röder, J.; Kübeck, C.; Wisotzky, F., 2017. Ex situ groundwater treatment
585 triggering the mobilization of geogenic uranium from aquifer sediments. – *Science of the To-
586 tal Environment* 587-588, 371-380.

587 Bayer, M., 1997. *Natürliche Arsenanreicherungen in der Oberen Süßwassermolasse Bayerns [Natural
588 arsenic accumulations in the Upper Süßwassermolasse of Bavaria]*. Veröffentlichungen des
589 Grundbauinstitutes der Landesgewerbeanstalt Bayern 77 (in German).

590 Bayer, M., Henken-Mellies, W.-U., 1998. *Untersuchung arsenführender Sedimente in der Oberen
591 Süßwassermolasse Bayerns [Investigation of arsenic bearing sediments of the Upper
592 Süßwassermolasse in Bavaria]*. *GeoCongress* 4, 42-58 (in German).

593 Berg, M., Tran, H.C., Nguyen, T.C., Pham, H.V., Schertenleib, R., Giger, W., 2001. Arsenic contami-
594 nation of groundwater and drinking water in Vietnam: a human health threat. *Environmental
595 Science and Technology* 35 (13), 2621-2626.

596 BfS (Bundesamt für Strahlenschutz, ed.), 2009. *Strahlenexposition durch natürliche Radionuklide im
597 Trinkwasser in der Bundesrepublik Deutschland [Radiation exposure by natural radionuclides
598 in drinking water in the Federal Republic of Germany]*. urn:nbn:de:0221-20100319945 (in
599 German).

600 Bhatia, M.R., Crook, K.A.W., 1986. Trace element characteristics of graywackes and tectonic setting
601 discrimination of sedimentary basins. *Contributions to Mineralogy and Petrology* 92, 181-193.

602 Birke, M., Rauch, U., Lorenz, H., Kringel, R., 2010. Distribution of uranium in German bottled and
603 tap water. *Journal of Geochemical Exploration* 107 (3), 272-282.

604 Bjørklund, G., Christophersen, O.A., Chirumbolo, S., Selinus, O., Aaseth, J. (2017). Recent aspects of
605 uranium toxicology in medical geology. *Environmental Research* 156, 526-533.

606 Bots, P., Behrends, T., 2008. Uranium mobility in subsurface aqueous systems: the influence of redox
607 conditions. *Mineralogical Magazine*, 72 (1), 381-384.

608 Brown, G.C., Locke, C.A., 1979. Space-time variations in British Caledonian granites – some geo-
609 physical correlations. *Earth and Planetary Science Letters* 45, 69-79.

610 Brown, C.J., Jurgens, B.C., Katz, B.G., Landon, M.K., Eberts, S.M., 2007. Arsenic and uranium in
611 four aquifer settings: Occurrence, distribution, and mechanisms for transport to supply wells.
612 NGWA Naturally Occurring Contaminants Conference: Arsenic, Radium, Radon, Uranium,
613 March 22-23, 2007, Charleston/U.S.A.

614 Bundschuh, J., Farias, B., Martin, R., Storniolo, A., Bhattacharya, P., Cortes, J., Bonorinom, G., Al-
615 bouy, R., 2004. Groundwater arsenic in the Chaco-Pampean Plain, Argentina: case study from
616 Robles country, Santiago del Estero Province. *Applied Geochemistry* 19 (2), 231-243.

617 Burkhalter, R.M., 1995. Ooidal ironstones and ferruginous microbialites: origin and relation to se-
618 quence stratigraphy (Aalenian and Bajocian, Swiss Jura mountains). *Sedimentology* 42, 57-74.

619 Condie, K.C., 1993. Chemical composition and evolution of the upper continental crust: Contrasting
620 results from surface samples and shales. *Chemical Geology* 104, 1-37.

621 Cremer, N., Obermann, P., Wisotzky, F., 2003. Mobilization of Nickel, Cobalt and Arsenic in a Multi-
622 aquifer Formation of the Lower Rhine Valley: Identification and Modeling of the Processes
623 Controlling Metal Mobility. In: Schulz, H.D., Haderer, A. (Eds.) *Geochemical Processes in*
624 *Soil and Groundwater – Measurement – Modelling – Upscaling* (Proceedings GeoProc2002),
625 3-18, Wiley, Weinheim.

626 Cullers, R.L., 1994. The chemical signature of source rocks in size fractions of Holocene stream sedi-
627 ment derived from metamorphic rocks in the Wet Mountains region, Colorado, U.S.A.
628 *Chemical Geology* 113, 327-343.

629 Cuney, M. (2010): Evolution of Uranium Fractionation Processes through Time: Driving the Secular
630 Variation of Uranium Deposit Types. – *Economic Geology* (2010) 105 (3): 553–569.

631 Del Razo, L.M., Arellano, M.A., Cebrián, M.E., 1990. The Oxidation States of Arsenic in Well-Water
632 from a Chronic Arsenicism Area of Northern Mexico. *Environmental Pollution* 64 (2), 143-
633 153.

634 Dill, H.G., Wehner, H., 1999. The depositional environment and mineralogical and chemical composi-
635 tions of high ash brown coal resting on early Tertiary saprock (Schirnding Coal Basin, SE
636 Germany). *International Journal of Coal Geology* 39, 301-328.

637 Doi, K., Hirono, S., Sakamaki, Y., 1975. Uranium Mineralization by Groundwater in Sedimentary
638 Rocks, Japan. *Economic Geology* 70, 628-646.

639 Foodwatch e.V., 2009. Uran-Werte im Trinkwasser (Stand Herbst 2009) [Uranium concentrations
640 in drinking water, status: autumn 2009]. [http://foodwatch.de/kampagnen___themen/mineralwas-](http://foodwatch.de/kampagnen___themen/mineralwasser/trinkwasser/index_ger.html)
641 [ser/trinkwasser/index_ger.html](http://foodwatch.de/kampagnen___themen/mineralwasser/trinkwasser/index_ger.html) (in German).

- 642 Frengstad, B., Skrede, A.K.M., Banks, D., Krog, J.R., Siewers, U., 2000. The chemistry of Norwegian
643 groundwaters: III. The distribution of trace elements in 476 crystalline bedrock groundwaters,
644 as analysed by ICP-MS techniques. *The Science of the Total Environment* 246, 21-40.
- 645 Giblin, A.M., Batts, B.D., Swaine, D.J., 1981. Laboratory simulation studies of uranium mobility in
646 natural waters. *Geochimica et Cosmochimica Acta* 45, 699-709.
- 647 Goldberg, G., Lepper, J., Röhling, H.G., 1995. Geogene Arsengehalte in Gesteinen und Grundwässern
648 des Buntsandsteins in Südniedersachsen [Geogenic arsenic contents in rocks and groundwater
649 of the Buntsandstein formation in southern Lower Saxony, Germany]. *Zeitschrift für
650 Angewandte Geologie* 41, 118-124 (in German).
- 651 Gromet, L.P., Dymek, R.F., Haskin, L.A., Korotev, R.L., 1984. The "North American Shale Compo-
652 site": Its compilation, major and trace element characteristics. *Geochimica et Cosmochimica
653 Acta* 48, 2469-2482.
- 654 Gwinner, M.P., 1971. *Geologie der Alpen – Stratigraphie, Paläogeographie, Tektonik* [Geology of the
655 Alps – stratigraphy, paleogeography, tectonics]. Schweizerbart, Stuttgart, 477 pp (in German).
- 656 Heinrichs, G., Udluft, P., 1996. Geogenes Arsen in Grundwässern Deutschlands unter
657 Berücksichtigung der Aquifergeologie [Geogenic arsenic in German groundwater with respect
658 to aquifer geology]. *Zeitschrift der Deutschen Geologischen Gesellschaft* 147, 519-530 (in
659 German).
- 660 Heinrichs, G., Udluft, P., 1999. Natural arsenic in Triassic rocks: A Source of drinking water contami-
661 nation in Bavaria, Germany. *Hydrogeology Journal* 7, 468-476.
- 662 Henningsen, D., Katzung, G., 2002. *Einführung in die Geologie Deutschlands* [Introduction to the
663 geology of Germany]. Spektrum, Berlin (in German).
- 664 Herch, A., 1997. Untersuchungen zur hydrogeochemischen Charakteristik der Spurenelemente und
665 Schwefelspezies im Aachener Thermalwasser [Study on hydrogeochemical characteristics of
666 trace elements and sulphur species in the thermal water of Aachen]. *Mitteilungen zur
667 Ingenieurgeologie und Hydrogeologie* 64, RWTH Aachen University (in German).
- 668 HLUG (Hessisches Landesamt für Umwelt und Geologie, ed.), 2008. Urankonzentrationen in
669 Hessischen Grundwässern – eine Studie des Hessischen Landesamtes für Geologie und
670 Umwelt [Uranium concentrations in Hessian groundwater – a study of the Hessian authority
671 for Geology and Environment] (in German).
- 672 Hofmann, B.A., 1990. Reduction spheroids from northern Switzerland – Mineralogy, geochemistry
673 and genetic models. *Chem. Geol.* 81, 55-81.
- 674 ICS (International Commission on Stratigraphy, ed.), 2009. International stratigraphic chart.
- 675 Katsoyiannis, I.A., Hug, S.J., Ammann, A., Zikoudi, A., Hatziliontos, C., 2007. Arsenic speciation and
676 uranium concentrations in drinking water supply wells in Northern Greece: Correlations with
677 redox indicative parameters and implications for groundwater treatment. *Science of the Total
678 Environment* 383 (1-3), 128-140.
- 679 Kipp, G.G., Stone, J.J., Stetler, L.D., 2009. Arsenic and uranium transport in sediments near aban-
680 doned uranium mines in Harding County, South Dakota. *Applied Geochemistry* 24, 2246-
681 2255.
- 682 Kneuper, G., List, K.-A., Maus, H., 1977. Geologie und Genese der Uranmineralisation des Oostroges
683 im Nordschwarzwald [Geology and genesis of the uranium mineralization of the Oostrog in
684 the northern Black Forest]. *Erzmetall* 30 (11), 522-525 (in German).

- 685 Krebs, W., 1976. The Tectonic Evolution of Variscan Meso-Europe, In: Ager, D.V., Brooks, M.
686 (eds.), Europe from Crust to Core. Wiley, London.
- 687 Kumar, A.; Karpe, R.K.; Rout, S.; Gautam, Y.P.; Mishra, M.K.; Ravi, P.M.; Tripathi, R.M. (2016):
688 Activity ratios of $^{234}\text{U}/^{238}\text{U}$ and $^{226}\text{Ra}/^{228}\text{Ra}$ for transport mechanisms of elevated uranium in
689 alluvial aquifers of groundwater in south-western (SW) Punjab, India. – Journal of
690 Environmental Radioactivity 151: 311-320.
- 691 Kumar, A.; Singh, C.K. (2020): Arsenic enrichment in groundwater and associated health risk in
692 Baridob region of Indus basin, Punjab, India. – Environmental Pollution 256, 1-13.
- 693 Kunkel, R., Voigt, H.-J., Wendland, F., Hannappel, S., 2004. Die natürliche, ubiquitär überprägte
694 Grundwasserbeschaffenheit in Deutschland [The natural, ubiquitarily overprinted,
695 groundwater quality in Germany]. Schriften des Forschungszentrums Jülich, Reihe
696 Umwelt/Environment 47, 204 pp, Jülich.
- 697 Kurth, D., 2010. Distribution and origin of geogenic uranium in groundwater systems of the southern
698 Lower Rhine Embayment and adjacent areas. B.Sc. thesis, RWTH Aachen University, Ger-
699 many (unpublished).
- 700 Lapworth, D.J.; Krishan, G.; MacDonald, A.M.; Rao, M.S. (2017): Groundwater quality in the alluvial
701 aquifer system of northwest India: new evidence of the extent of anthropogenic and geogenic
702 contamination. – Science of the Total Environment 599-600: 1433-1444.
- 703 LGL, LfU (Bayerisches Landesamt für Gesundheit und Lebensmittelsicherheit, Bayerisches
704 Landesamt für Umwelt, Eds.), 2007. Untersuchungen zum Vorkommen von Uran im Grund-
705 und Trinkwasser in Bayern [Study of uranium occurrence in groundwater and drinking water
706 in Bavaria] (in German).
- 707 Lowers, H.A., Breit, G.N., Foster, A.L., Whitney, J., Yount, J., Uddin, M.N., Muneem, A.A., 2007.
708 Arsenic incorporation into authigenic pyrite, Bengal Basin sediment, Bangladesh. Geochimica
709 et Cosmochimica Acta 71, 2699-2717.
- 710 Matschullat, J., 1999. Arsen in der Geosphäre [Arsenic in the geosphere]. In: Rosenberg, F., Röhling,
711 H.G. (Eds.), Arsen in der Geosphäre. Schriftenreihe der Deutschen Geologischen Gesellschaft
712 6, 5-20 (in German).
- 713 Merkel, B., Sperling, B., 1998. Hydrogeochemische Stoffsysteme, Teil II [Hydrogeochemical
714 substances, part II]. Schriftenreihe des Deutschen Verbandes für Wasserwirtschaft und
715 Kulturbau (DVWK) 117, Bonn (in German).
- 716 Mertens, J., 2000. Anreicherung und Bindungsformen geogener Arsen- und Schwermetall-belastungen
717 glaukonitreicher Kreidesedimente im mittleren Ruhrgebiet [Enrichment and fractionation of
718 geogenic arsenic and heavy metals in glauconite-rich Cretaceous sediments in the Middle
719 Ruhr Area.]. M.Sc. thesis, University of Essen, Germany (in German).
- 720 Meurer, M.; Banning, A., 2019. Uranmobilisierung im Helgoländer Buntsandstein – Auswirkungen
721 auf die Brack- und Trinkwasserqualität. – Grundwasser 24, 43-50.
- 722 Missana T., García-Gutiérrez, M., Maffiotte, C., 2003. Experimental and modelling study of the urani-
723 um(VI) sorption on goethite. Journal of Colloid and Interface Science 260, 291-301.
- 724 Nicolli, H.B., Bundschuh, J., García, J.W., Falcón, C.M., Jean, J.-S., 2010. Sources and controls for
725 the mobility of arsenic in oxidizing groundwaters from loess-type sediments in arid/semi-arid
726 dry climates – Evidence from the Chaco-Pampean plain (Argentina). Water Research 44 (19),
727 5589-5604.

- Orloff, K.G., Mistry, K., Charp, P., Metcalf, S., Marino, R., Shelly, T., Melaro, E., Donohoe, A.M., Jones, R.L., 2004. Human exposure to uranium in groundwater. *Environmental Research* 94, 319-326.
- Oyarzun, R., Lillo, J., Higuera, P., Oyarzun, J., Maturana, H., 2004. Strong arsenic enrichment in sediments from the Elqui watershed, Northern Chile: industrial (gold mining at El Indio-Tambo district) vs. geologic processes. *Journal of Geochemical Exploration* 84, 53-64.
- Patnaik, R.; Lahiri, S.; Chahar, V.; Naskar, N.; Sharma, P.K.; Avhad, D.K.; Bassan, M.K.T.; Knolle, F.; Schnug, E.; Srivastava, A., 2015. Study of uranium mobilization from Himalayan Siwaliks to the Malwa region of Punjab state in India. – *Journal of Radioanalytical and Nuclear Chemistry* 308: 913-918.
- Pedersen, H.D., Postma, D., Jakobsen, R., 2006. Release of arsenic associated with the reduction and transformation of iron oxides. *Geochimica et Cosmochimica Acta* 70, 4116-4129.
- Planer-Friedrich, B., Armienta, M.A., Merkel, B.J., 2001. Origin of arsenic in the groundwater of the Rioverde basin, Mexico. *Environmental Geology* 40, 1290-1298.
- Plinninger, R.J., Thuro, K., 1999. Die geologischen Verhältnisse beim Vortrieb des Meisterntunnels Bad Wildbad/Nordschwarzwald [Geological conditions during construction of the Meistertunnel Bad Wildbad/northern Black Forest]. *Jahrbuch 1999 des Oberrheinischen Geologischen Vereins*, 1-11 (in German).
- Ravenscroft P., Burgess W.G., Ahmed K., Burren M., Perrin J., 2005. Arsenic in groundwater of the Bengal Basin, Bangladesh: Distribution, field relations, and hydrogeological setting. *Hydrogeology Journal* 13, 727-751.
- Rosenberg, F., Mittelbach, G., Kirnbauer, T., 1999. Geogene Arsengehalte im Bereich der Wiesbadener Thermalquellen [Geogenic arsenic contents within the range of the thermal springs of Wiesbaden]. In: Rosenberg, F., Röhling, H.-G. (eds.): *Arsen in der Geosphäre, Schriftenreihe der Deutschen Geologischen Gesellschaft* 6, 101-106 (in German).
- Rüde, T.R., 1996. Beiträge zur Geochemie des Arsens [Contributions to arsenic geochemistry]. *Karlsruher Geochemische Hefte* 10, 1-206 (in German).
- Sauer, K., Simon, P., 1975. Die Eisenerze des Aalenium und Bajocium im Oberrheingraben (Grube Kahlenberg, Grube Schönberg und kleinere Vorkommen) [Aalenian and Bajocian iron ores in the Upper Rhine Graben (Kahlenberg quarry, Schönberg quarry and smaller deposits)]. *Geologisches Jahrbuch D10*, 25-68 (in German).
- Scheid, Y., Schiedek, T., Ebhardt, G., 1999. Geogenes Arsen in Quellwässern des Spessarts (Bayern) [Geogenic arsenic in spring waters of the Spessart (Bavaria)]. *Schriftenreihe der Deutschen Geologischen Gesellschaft* 6, 127-130 (in German).
- Schindler, F., Wisotzky, F., Banning, A., Cremer, N. (2016). Uranmobilisierung im Buntsandstein des Mechnicher Triasdreiecks [Uranium mobilization from Buntsandstein sediments of the Mechnic Triassic Triangle]. In: Blum, P., Goldscheider, N., Göppert, N., Kaufmann-Knoke, R., Klinger, J., Liesch, T., Stober, I. (eds.): *Grundwasser – Mensch – Ökosysteme*: 262-263. doi: 10.5445/KSP/1000051730 (in German).
- Sherman, H.M., Gierke, J.S., Anderson, C.P., 2007. Controls on spatial variability of uranium in sandstone aquifers. *Ground Water Monitoring & Remediation* 27 (2), 106-118.
- Smedley, P.L., Kinniburgh, D.G., 2002. A review of the source, behaviour and distribution of arsenic in natural waters. *Applied Geochemistry* 17, 517-568.

- 771 Smedley, P.L., Smith, B., Abesser, C., Lapworth, D., 2006. Uranium occurrence and behaviour in
772 British groundwater. British Geological Survey Commissioned Report, CR/06/050N.
- 773 Spoljaric, N., Crawford, W.A., 1978. Glauconitic Greensand: A possible filter of heavy metal cations
774 from polluted water. *Environmental Geology*, 2 (4), 215-221.
- 775 Stanger, G., 2005. A palaeo-hydrogeological model for arsenic contamination in southern and south-
776 east Asia. *Environmental Geochemistry and Health* 27, 359-367.
- 777 Steffanowski, J.; Banning, A. (2017): Uraniferous dolomite – a natural source of high groundwater
778 uranium concentrations in northern Bavaria, Germany? – *Environmental Earth Sciences* 76,
779 508-518.
- 780 Stewart, B.D., Mayes, M.A., Fendorf, S. (2010): Impact of Uranyl-Calcium-Carbonato Complexes on
781 Uranium(VI) Adsorption to Synthetic and Natural Sediments. *Environmental Science &*
782 *Technology* 44, 928-934.
- 783 Stollenwerk, K.G., 2002. Geochemical Processes Controlling Transport of Arsenic in Groundwater: A
784 Review of Adsorption. In: Welch, A.H. and Stollenwerk, K.G. (Eds.) *Arsenic in Ground Wa-*
785 *ter – Geochemistry and Occurrence*, 67-100, Springer, New York.
- 786 Taylor, S.R., 1964. Abundance of chemical elements in the continental crust: a new table. *Geochimica*
787 *et Cosmochimica Acta* 28, 1273-1285.
- 788 Unger, H.J., 1989. Die Lithozonen der Oberen Süßwassermolasse Südostbayerns und ihre vermutlich
789 zeitlichen Äquivalente gegen Westen und Osten [Lithozones of the Upper Süßwassermolasse
790 of southeastern Bavaria and its assumed equivalents towards the west and the east]. *Geologica*
791 *Bavarica* 94, 195-223 (in German).
- 792 Van Berk, W., Fu, Y. (2017). Redox Roll-Front Mobilization of Geogenic Uranium by Nitrate Input
793 into Aquifers: Risks for Groundwater Resources. *Environmental Science & Technology* 51,
794 337-345.
- 795 Wagner, J.-F., 1999. Arsen im Grundwasser der triadischen Randfazies Luxemburgs [Arsenic in the
796 triadic marginal facies of Luxembourg]. *Schriftenreihe der Deutschen Geologischen*
797 *Gesellschaft* 6, 131-132 (in German).
- 798 Walenta, K., Hatert, F., Theye, T., Lissner, F., Röller, K., 2009. Nielsbohrite, a new potassium uranyl
799 arsenate from the uranium deposit of Menzenschwand, southern Black Forest, Germany. *Eu-*
800 *ropean Journal of Mineralogy* 21, 515-520.
- 801 Welch, A.H., Lico, M.S., 1998. Factors controlling As and U in shallow ground water, southern Car-
802 son Desert, Nevada. *Applied Geochemistry* 13 (4), 521-539.
- 803 Wendland, A., Rank, G., Barth, A., 1999. Verteilung des Arsens in den Rotliegendensedimenten
804 Sachsens [Arsenic distribution in Rotliegend sediments of Saxony]. *Schriftenreihe der*
805 *Deutschen Geologischen Gesellschaft* 6, 93-100 (in German).
- 806 Winkelmann, I., Thomas, M., Vogl, K., 2001. Aerial measurements on uranium ore mining, milling
807 and processing areas in Germany. *Journal of Environmental Radioactivity* 53, 301-311.
- 808 Wolkersdorfer, C., 1996. Hydrogeochemical investigations of an abandoned uranium mine in the Erz-
809 gebirge/Germany. *Applied Geochemistry* 11, 237-241.
- 810 Yudovich, Y.E., Ketris, M.P., 2005. Arsenic in coal: a review. *International Journal of Coal Geology*
811 61, 141-196.
- 812 Zahid, A., Hassan, M.Q., Breit, G.N., Balke, K.-D., Flegr, M., 2009. Accumulation of iron and arsenic
813 in the Chandina alluvium of the lower delta plain, Southeastern Bangladesh. *Environmental*
814 *Geochemistry and Health* 31, 69-84.

- 815 Zamora, M.L., Tracy, B.L., Zielinski, J.M., Meyerhof, D.P., Moss, M.A., 1998. Chronic Ingestion of
816 uranium in Drinking Water: A Study of Kidney Bioeffects in Humans. *Toxicological Sciences*
817 43, 68-77.
- 818 Ziegler, P.A., 1990. *Geological Atlas of Western and Central Europe*. Elsevier, Amsterdam.